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Response of free-living Nematodes to the quality of water and sediment at Bou Chrara Lagoon (Tunisia) during winter 2000

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Abstract

The Bou Ghrara lagoon, a stretch of water in southeastern Tunisia, has shown an alarming reduction of its fishery resources since 1993. In order to study the response of free-living nematodes to the water and sediment quality of this area, thirteen stations have been sampled. According to this study, the heavy metal, organic carbon and hydrocarbon content of sediments are key factors negatively influencing the density, biomass and diversity of the nematofauna.

Keywords: Lagoon, Free-living nematodes, Diversity, Hydrocarbon, Heavy metal.

Introduction

The Bou Ghrara lagoon with an area of 500 km² is the biggest of all Tunisian lagoons. This important ecosystem has experienced an alarming reduction of its fishery resources since 1993. Before presenting the results of this work, it seems necessary to stress the importance of the meiofauna, one of the first links in the food web (SOGARD, 1984; CASTEL, 1992; AMARA & BODIN, 1995).

The meiobenthos and notably the freeliving nematodes, the most abundant taxa (e.g. TEAL & WIESER, 1966; ALKEMADE *et al.*, 1993; BEYREM & AISSA, 2000), have been used in the impact studies of many aquatic ecosystems and several authors have demonstrated that these animals are ideal for pollution monitoring because of their potential for rapid response to environmental changes (e.g. WARWICK, 1993; BEYREM & AISSA, 2000, BURTON *et al.*, 2001, MAHMOUDI *et al.*, 2002).

It is no longer necessary to demonstrate the relevance of meiofauna as indicators of environmental stress. Indeed, ANSARI *et al.* (1984) noted a significant spatial and temporal variation in mean meiofauna density in Goa (India), in response to organic discharge via sewage and prevailing environmental conditions. SCHRATZBERGER & WARWICK (1998) revealed in a microcosm experiment that total abundance, diversity and species richness of the nematofauna decreased significantly with increasing levels of organic enrichment. BEYREM & AISSA (2000) showed, in the Bay of Bizerta (Tunisia), that free-living nematode biomass and diversity were altered by hydrocarbon pollution. At Ghar El Melh lagoon (Tunisia), MAHMOUDI *et al.* (2002) reported that the salinity and dissolved oxygen of waters and the ammonia content of sediments affected density, biomass and the diversity of the nematofauna.

The aims of this study were to investigate the response of free-living nematodes to the quality of waters and sediments at Bou Ghrara lagoon and to validate the use of meiofauna in impact studies of lagoons.

Materials and Methods

Sampling and nematofauna analyses

Sediments were collected from 13 sampling stations in Bou Ghrara lagoon in February

2000 (Fig. 1). The stations' depth ranged from 0.9-10 m. For meiofauna analysis, three replicate 10 cm² hand-cores were taken at each station and fixed immediately in 4% buffered formaldehyde. Sampling depth was limited to the first 18 cm of the substrate. Meiofaunal taxa, defined here as metazoans that pass through a 1 mm mesh sieve and are retained on a 0.040mm sieve (VITIELLO & DINET, 1979), were sieved following the resuspensiondecantation methodology (WIESER, 1960) and stained with Rose-Bengal (0.2 g/l). Nematodes were counted under a microscope and 100 randomly picked individuals (KOTTA & BOUCHER, 2001) were identified to genus or species using the pictorial keys of PLATT & WARWICK (1983, 1988), WARWICK et al. (1997) and AISSA & HERMI (1997). These individuals were also used for evaluating individual and total biomass. Their lengths and widths were measured (VITIELLO & AISSA,



Fig. 1: Geographical location of the stations sampled in Bou Ghrara lagoon during winter 2000 (BG1 - BG13).

1985). The body volume of nematodes was estimated using the andrassy formula $V = LW^2$ / 16x 10⁵ (ANDRASSY, 1956), where V is the volume in nanolitres, L the length and W the maximum width (L and W expressed in micrometres). The wet weight of each nematode was obtained from multiplying volume by specific gravity (assumed to be 1.13) (FELLER & WARWICK, 1988). Dry weight was assumed to be 25% of the wet weight (JUARIO, 1975).

Hydrological parameters

Water temperature, dissolved oxygen and salinity were measured *in situ* by using a field thermometer-oxymeter (CG867) and a field salinometer (WTW LF 196).

Sediment parameters

For sediment parameter analysis, one replicate 10 cm² hand-core was used. Grain size distribution for each station was measured by the wet sieve method of COULL (1970) and mean grain size (*MZ*) and sorting coefficient ($\% \circ \Delta I$) (TRASK, 1932) were calculated. Hydrocarbon content in sediment was quantified spectrophotometrically (KANTIN, 1977) following the protocol described by BEYREM & AISSA (1998). Total organic carbon (*TOC*) was determined coulometrically using a Coulomat 702. Heavy metal analysis was performed by atomic absorption spectrometer.

Data processing

Nematode abundance, mean individual weight, mean total biomass, diversity (Shannon-Weaver index H') (SHANNON & WEAVER, 1963) and evenness (Pielou's J) (PIELOU, 1969) were calculated for each station.

Species abundance data were presented in k-dominance plots, in which species are ranked in decreasing order of dominance. The percentage cumulative abundance (k-dominance) is then plotted against the species rank k (LAMBSHEAD *et al.*, 1983, PLATT *et*

al., 1984; SCHRATZBERGER & WARWICK, 1998).

Data were tested for homogeneity of variance and for normal distribution. The following statistical treatment of the results consisted of pairwise comparisons (Student ttest) which allowed evaluating the influence of sampling stations on nematode distribution. Pearson correlations (r) (SCHWARTZ, 1983) were used to test for relationships between environmental variables and nematode density, biomass and diversity. Faunal affinity between stations was measured by regrouping the stations on the basis of Euclidean distance (LEGENDRE & LEGENDRE, 1998). The raw data were transformed to $\log_{10} (x + 1)$ in order to stabilise the variance and all tests were performed using the SYSTAT 8.0 software package (WILKINSON et al., 1992) and STATBOXTM 2.5 for Windows (GRIMMER Logiciels).

Results

Hydrological parameters

Water temperature and salinity did not show any appreciable change at the thirteen sampling sites. However, water dissolved oxygen content ranged from 7.68 mg l^{-1} at station (BG9) to 11.60 mg l^{-1} at site (BG6) (Fig. 2).

Sediment parameters

Table 1 gives data on the physical and chemical environmental factors of the sediment. Granulometric analysis showed that the sediment was mainly composed of larger to fine sands which were well sorted at all stations. Sediment hydrocarbon concentration fluctuated between $0.05 - 4.80 \text{ mg g}^{-1}$ dry weight (*dw*) and the lowest values were recorded at stations near the marine inputs (BG6, BG 9 and BG13). Sediment total organic carbon content (*TOC*) ranged from 0.07 to 7.56 % (*dw*) and the lowest values are also observed at sites BG6, BG9 and BG13. Heavy metal (Zn, Cr, Pb) concentrations in sediment were lower than international norms for polluted marine sediments (RAPPORT DGPA, 2000).

Nematofauna

Nematofaunal abundance ranged from 154 \pm 7.81 individuals 10 cm⁻² at station BG2 to 5641 \pm 28.05 individuals 10 cm⁻² (mean \pm SE)

at site BG13 (Fig. 3). Highest densities were recorded at stations near the marine inputs (BG6, BG9 et BG13) and the lowest densities were found at the most isolated sites BG1 and BG2. Student t-test between stations supported the existence of a nematode density gradient from BG1 and BG2 to BG6, BG9 et BG13 and indicated that Bou Ghrara lagoon is an heterogeneous area (Table 2).

Table 1

Positions (latitude and longitude), depth and data on sediment parameters of the 13 stations investigated in Bou Ghrara lagoon during winter 2000. Hs: Sediment hydrocarbon content in mg g¹ dry weight, *TOC*: Sediment total organic carbon content in % dry weight, Pb: Sediment Pb content in ppm , Zn: Sediment Zn content in ppm, Cr: Sediment Cr content in ppm, FF: Sediment fine fraction content in %, MZ: mean grain size in μm , ΔI : sorting coefficient.

Stations	Latitude	Longitude	Depth	Hs	TOC	Pb	Zn	Cr	FF	MZ	‰∆I
	(South)	(East)	(m)	(mg g ⁻¹)	(%)	(ppm)	(ppm)	(ppm)	(%)	(µm)	
BG1	33° 32' 23	10° 42' 03	10	3.42	4.8	63	324	30	0.80	140	1.44
BG2	33° 30' 56	10° 43' 87	9.5	1.99	3.31	72	475	42	67.60	980	1.32
BG3	33° 30' 56	10° 43' 87	1.2	2.94	4.65	22	15	15	47.80	320	2.14
BG4	33° 32' 39	10° 46' 25	3.8	1.57	1.21	50	13	7	4.00	250	2.15
BG5	37° 30' 34	10° 48' 47	2.5	6.4	5	30	14	14	33.00	400	0.40
BG6	33° 37' 88	10° 44' 64	1.4	0.05	0.07	21	19	13	7.20	130	1.41
BG7	33° 35' 88	10° 46'80	15.9	2.87	7.04	29	649	38	44.80	170	1.60
BG8	33° 32' 17	10° 50' 87	4.2	1.67	1.66	24	15	11	16.20	140	1.44
BG9	33° 41' 70	10° 44' 85	10.9	0.07	0.02	11	12	9	97.60	650	0.32
BG10	37° 08' 47	10° 11' 33	10.2	4.8	7.56	41	32	28	97.40	210	0.34
BG11	33° 41' 55	10° 49' 42	4.1	1.84	2.16	24	13	6	14.20	160	1.25
BG12	33° 39' 50	10° 51' 56	3.2	0.63	0.5	12	10	8	0.60	190	1.37
BG13	37° 37' 77	10° 53' 54	0.9	0.7	0.83	12	12	6	77.80	90	0.25



Fig. 2: Spatial variation of some hydrological parameters values in Bou Ghrara during winter 2000. O2: water dissolved oxygen content (mg l⁻¹), Sal. : Salinity (g l⁻¹), T°: Water temperature (°C). The stations 1-13 correspond to BG1 - BG13 in Fig.1.

Nematode mean individual weight fluctuated between $0.19 \,\mu g (dw)$ at station BG1 to $0.81 \,\mu g (dw)$ at station BG13 (Fig. 4). Following the same trend as observed for abundance, the highest and lowest mean individual weight values were found at stations near the marine inputs (BG6, BG9 et BG13) and at the most isolated sites (BG1 and BG2), respectively.

Nematofaunal mean total biomass ranged from $33.82 \ \mu\text{g} \ 10 \ \text{cm}^{-2} \ (dw)$ at station BG1 to

 $4569.21 \ \mu g \ 10 \ cm^{-2} (dw)$ at site BG13. The nematode biomass showed three peaks corresponding to marine influenced stations (Fig. 4).

A total of 47 free-living nematode species were identified. A list of these species arranged in decreasing mean general dominance (MGD) is presented in Table 3. Of the 47 species identified *Terschellingia longicaudata* was the most abundant form. This species was followed by *Paracomesoma dubium*,



Fig. 3: Spatial variation of nematode mean densities (in individuals 10 cm⁻²) in Bou Ghrara lagoon during winter 2000. The stations 1-13 correspond to BG1- BG13 in Fig.1.

Table 2
Significance of differences between nematofaunal abundances as a function of sampling stations.

Stations	BG1	BG2	BG3	BG4	BG5	BG6	BG7	BG8	BG9	BG10	BG11	BG12	BG13
BG1		1,67	17,54	16,35	13,38	216,90	20,42	17,45	17,80	64,49	76,95	49,85	323,88
BG2	NS		22,76	20,89	22,55	218,99	24,94	19,65	73,13	69,12	81,25	51,08	326,46
BG3	***	***		0,87	1,91	216,99	1,13	9,90	68,44	61,05	74,65	46,18	327,19
BG4	***	***	NS		1,70	214,06	1,06	9,26	67,64	55,21	69,03	45,61	322,58
BG5	***	***	NS	NS		210,43	0,98	7,14	66,04	48,043	62,44	44,24	317,46
BG6	***	***	***	***	***		211,72	179,88	87,94	187,66	176,33	101,80	71,76
BG7	***	***	NS	NS	NS	***		6,74	66,11	49,92	64,13	44,19	319,75
BG8	***	***	**	**	* *	***	***		56,16	21,95	32,86	36,75	271,03
BG9	***	***	***	***	***	***	***	* * *		48,68	41,16	14,48	154,16
BG10	***	***	***	***	***	***	***	* * *	***		17,25	27,68	292,06
BG11	***	***	***	***	***	***	***	* * *	***	* * *		20,66	278,34
BG12	***	***	***	***	***	***	* * *	***	***	* * *	* * *		166,80
BG13	***	***	***	***	***	***	* * *	***	***	* * *	* * *	* * *	

NS: no significant difference at p < 0.05,

**: significant difference at p < 0.01,

***: very significant difference at p < 0.001.



Fig. 4: Spatial variation of total biomasses (BT in dry weight, μ g 10 cm⁻²) and mean individual weight (bi in dry weight, μ g) of nematodes in Bou Ghrara lagoon during winter 2000. The stations 1-13 correspond to BG1- BG13 in Figure 1.

Table 3
List of free-living nematode species from Bou Ghrara lagoon (February 2000). The list includes
only species whose dominance (D%) is higher than 1 $\%$ at least in one of the samples.
The dominance of the two most abundant species from each station is indicated in bold.

Stations	BG1	BG2	BG3	BG4	BG5	BG6	BG7	BG8	BG9	BG10	BG11	BG12	BG13	
Nematode	D %	D %	D %	D%	D%	D %	D%	D%	D %	D %	D %	D%	D%	DGM %
species														
Terschellingia	22.00	10.00	5.00	7.84	2.73	24.29	-	2.91	-	-	60.20	68.32	22.53	23.35
longicaudata														
De Man. 1907														
Paracomesoma	1.00	2.00	3.00	1.96	1.82	14.56	79.83	29.13	0.83	3.64	-	0.99	11.28	9.81
dubium														
(Filipjev. 1918)														
Daptonema fallax	-	-	-	62.76	1.82	9.71	3.85	1.94	-	-	-	1.98	0.75	7.76
Lorenzen. 1971														
Daptonema	64.00	70.00	75.00	5.88	2.73	-	-	38.84	-	61.80	13.27	5.94	0.75	7.76
normandicum														
De Man. 1890														
Metalinhomoeus	-	-	1.00	0.98	2.73	4.85	0.96	0.97	13.33	-	13.27	0.99	7.52	6.21
numidicus														
Aïssa et Vitiello. 1977														
Prochromadorella	-	1.00	-	-	-	-	-	1.94	32.54	-	-	-	3.76	4.84
neapolitana														
(De Man. 1876)														
Desmolaimus	-	-	-	-	-	-	1.92	-	3.33	-	-	-	11.24	4.44
demani														
Schulz. 1932														
Sabatieria granifer	-	-	-	-	8.18	-	-	-	-	-	2.04	1.98	10.53	4.19
Wieser. 1954														
Sabatieria	-	-	-	-	-	8.74	1.92	-	-	-	-	-	2.26	3.05
longicaudata														
Filipjev .1922														

Table 3 (Continued)

Terschellingia	8.00	_	3.00	_	1.82	4 85	1.92	_	_	3.64	-	1.98	2.26	2 56
communis	0.00	_	5.00	-	1.02	1.05	1.72	_		5.04	_	1.70	2.20	2.50
Do Mon 1999														
De Mail. 1000			1.00		70 71			20.20		1.55				2.00
Meioncholaimus	-	-	1.00	-	/4./1	-	-	20.39	-	4.55	-	-	-	2.00
pristiurus														
Filipjev.1918		0.00												
Terschellingia sp.	-	9.00	-	3.92	-	-	-	-	5.83	-	-	-	3.01	1.85
Microlaimus	-	-	-	-	-	-	-	-	-	12.73	-	7.92	1.5	1.69
cyatholaimoïdes														
De Man. 1922														
Viscosia glabra	-	-	-	-	-	-	-	-	5.83	4.55	-	-	2.26	1.61
Bastian. 1865														
Sabatieria pulchra	5.00	4.00	-	10.78	-	4.85	-	-	-	-	-	-	-	1.51
Riemann. 1970														
Halichoanolaimus	-	-	-	-	-	5.83	-	-	-	-	-	-	-	1.47
dolichurus														
Ssweljev. 1912														
Marilvnia	-	-	-	-	-	-	0.96	-	-	-	-	-	3.01	1.10
stekhoveni														
Wieser 1954														
Paralinhomoeus			-		_	3.88	1.92	_		_	-			1.02
tonuicaudatus	_	_	_	_	_	5.00	1.72	_		_	_	_		1.02
(Dütaabli 1974)														
(Butschill, 1674)			1.00			200								1.00
Prochromadoreua	-	-	1.00	-	-	3.88	-	-	-	-	-	-	-	1.00
longicauaata														
Lorenzen. 1972						2.00								0.00
Prooncholaimus	-	-	-	-	-	3.88	-	-	-	-	-	-	-	0.98
megastoma														
(Eberth. 1863)														
Chromadorita tenuis	-	-	2.00	1.96	-	-	0.96	-	5.00	3.64	-	1.98	-	0.92
(Schneider.1906)														
Sphaerolaimus	-	-	-	-	-	-	-	-	0.83	-	-	-	2.26	0.90
gracilis														
De Man. 1876														
Sabatieria punctata	-	-	-	-	0.91	-	-	-	-	-	-	-	2.26	0.83
Stekhoven. 1935														
Anticoma acuminata	-	-	-	-	-	-	-	-	-	-	-	-	2.26	0.81
(Eberth.1863)														
Desmodora ovigera	-	-	-	-	-	-	-	-	-	-	-	-	2.26	0.81
Ott. 1979														
Viscosia viscosa	-	-	-	-	-	-	-	-	-	-	-	-	2.26	0.81
(Bastian, 1865)														
Desmodora minuta	-	-	-	-	-	-	-	-	7.50	-	-	-	-	0.79
Wieser 1954									/					0172
Dorvlaimonsis	-	-	-	2 94	-	2.91	-	-	-	-	-	-	-	0.79
maditarranaus				2.74		2.71								0.75
De Zio 1068														
Chromaspirina				0.08					5.02					0.62
Chromaspirina	-	-	-	0.90	-	-	-	-	5.05	-	-	-	-	0.05
Filiniay 1019														
Filipjev .1918									5.00					0.(2
Camacolaimus	-	-	-	-	-	-	-	-	5.00	-	-	-	-	0.62
tardus														
De Man. 1989								1						

Table 3 (Continued)

Chain ania milana					0.01		0.06				2.06	1.00	0.75	0.62
Cobb. 1914	-	-	-	-	0.91	-	0.96	-	-	-	3.00	1.98	0.75	0.62
Metalinhomoeus	-	-	-	-	-	-	-	-	-	-	6.12	2.97	-	0.54
torosus														
Jensen et Gerlach, 1976														
Monhystera parva	-	-	-	-	-	-	-	-	-	-	-	-	15	0.54
(Bastian 1865)													1.0	0.01
Stylotheristus	-			_		7 77	0.96	-	_	_		-	15	0.54
mutilus		_		_	_	1.11	0.90	_	_	_	_	-	1.5	0.54
Loronzon 1072														
Lorenzen.1975													1.5	0.54
V asosioma sp.	-	-	-	-	-	-	-	-	-	-	-	-	1.3	0.54
Longicyatholaimus	-	-	-	-	-	-	-	-	5.00	-	-	-	-	0.55
longicaudatus														
(Marion. 1870)														
Oncholaimus	-	-	2.00	-	0.91	-	-	1.94	-	-	-	-	0.75	0.37
campylocercoïdes														
(De Coninck														
et Stekhoven. 1933)														
Filoncholaimus	-	-	-	-	-	-	-	-	3.33	-	-	-	-	0.36
ditlevseni														
(Kreis. 1932)														
Mesacanthion	-	-	-	-	-	-	-	-	2.50	-	-	-	-	0.26
hirsutum														
Gerlach. 1953														
Aponema	-	-	-	-	-	-	-	-	-	5.45	-	-	-	0.22
torosa														
(Lorenzen, 1973)														
Chromadorina	-	-	-	-	-	-	-	-	-	-	-	1.98	-	0.16
metulata												1.70		0110
Aïssa et Vitiello 1977														
Cvatholaimus prinzi	-	-	-	-		-		1 94	-	-	2.04			0.15
(Marion 1870)								1.71			2.01			0.15
Synonchiella eday	-		3.00		2 73	_	0.96	-	_	-		-	_	0.12
Aïssa et Vitiello, 1077		_	5.00		2.15	_	0.90			_	_	-	_	0.12
Naochromadora												0.00		0.08
neochiomuuonu	-	-	-	-	-	-	-	-	-	-	-	0.33	-	0.00
(Eilinion 1019)														
(Filipjev. 1916)			2.00											0.05
Euchromaaora	-	-	5.00	-	-	-	-	-	-	-	-	-	-	0.05
striata														
Eberth. 1803							1.02							0.04
Daptonema	-	-	-	-	-	-	1.92	-	-	-	-	-	-	0.04
proprium														
(Lorenzen. 1971)		• • •												0.04
Odontophora villoti	-	3.00	-	-	-	-	-	-	-	-	-	-	-	0.03
Luc et de Coninck. 1959														
Theristus flevensis	-	-	1.00	-	-	-	-	-	-	-	-	-	-	0.02
Stekhoven. 1935														
Calomicrolaimus	-	1.00	-	-	-	-	-	-	-	-	-	-	-	0.01
honestus														
(Jayasree et														
Warwick, 1970)														

Daptonema fallax, Daptonema normandicum, Metalinhomoeus numidicus, Prochromadorella neapolitana, Desmolaimus demani, Sabatieria granifer, Sabatieria longicaudata, Terschellingia communis and Metoncholaimus pristiurus. These 11 species comprise more than 75% of the total number of nematodes collected.

As shown in Fig. 5, diversity (H') and evenness (J) were greatest at stations near the marine inputs (BG6, BG9 et BG13). The most isolated sites (BG1 and BG2), dominated by *Daptonema normandicum*, had the lowest nematode diversity and evenness. This species was absent from stations BG6, BG9 and BG13 (Table 3).

The overlapping of the k-dominance curves of the stations makes it difficult to compare their nematode community diversity. However, k-dominance curves of the sites near the marine inputs (BG6, BG9 et BG13) show a clear picture of decreasing dominance and increasing diversity (Fig. 6).

isolated sites (BG1 and BG2), dominated by The regrouping of the sampling sites on *Daptonema normandicum*, had the lowest the basis of Euclidean distance according to



Fig. 5: Spatial variation of diversity indices (H' in bits and E) in Bou Ghrara lagoon during winter 2000. The stations 1-13 correspond to BG1-BG13 in Figure 1.



Fig. 6: Spatial variation of k-dominance curves in Bou Ghrara lagoon during winter 2000.

the Shannon diversity index values (H' in bits) showed a nematode diversity gradient from BG2 to BG6, BG9 and BG13 (Fig.7).

Discussion

Sediment hydrocarbon content recorded in Bou Ghrara lagoon, an ecosystem that has important fisheries, were comparable to those found in polluted marine sediment (Table 4).

The *TOC* concentrations we measured in Bou Ghrara lagoon were comparable to those observed in polluted sediments of the south lagoon (0.75-5.36% dw) (HERMI, 2001) and higher than those measured in sediments from the Aegean Sea (0.22 - 0.78% dw)(HATZIANESTIS *et al.*, 1998).

The total nematode abundance, mean individual weight, mean total biomass and diversity that we recorded were comparable to those reported in polluted marine sediments (Table 5).

There are several environmental variables that determine the distribution of interstitial fauna (SWEDMARK, 1964; McINTYRE, 1971). In Bou Ghrara lagoon, the organic carbon, hydrocarbon and heavy metal content



Fig. 7: Regrouping of the stations sampled in Bou Ghrara lagoon during winter 2000 on the basis of Euclidean distance according to their Shannon diversity indices (H' in bits).

 Table 4

 Bibliographic data on hydrocarbon contents reported from sediments of western Mediterranean.

Biotope	Sediment hydrocarbon content	Reference
Algerian coast	0.001- 36 µg g ⁻¹	Sellali et al., 1993
Tunis lagoon	16.5 - 150 μg g ⁻¹	Ghabi et Hadj Ali, 1993
Gulf of Tunis	1.8 - 122.5 μg g ⁻¹	Ghabi et Hadj Ali, 1993
Bizerta lagoon	2 - 20 μg g ⁻¹	Ghabi et Hadj Ali, 1993
Gulf of Fos/mer	0.04 - 47.58 g kg ⁻¹	Le Dréau et al., 1997
Bay of Bizerta (1992)	0.07 - 7.61 mg g ⁻¹	Beyrem et Aïssa, 1998
Bay of Bizerta (1996)	2.37 - 3.39 mg g ⁻¹	Beyrem et Aïssa, 2000
Bou Ghrara lagoon	$0.05 - 6.4 \text{ mg g}^{-1}$	<i>This work</i>

Table 5

Sites	Abundance	Individual weight	Total biomass	H'	Reference
	Ind. 10 cm ⁻²	(µg dw)	(µg 10 cm ⁻² dw)		
Marseilles'	2 - 4218	0.07* - 0.24*	0.28* - 844.5*	0.7 -7.34	Keller, 1984
sewage outfall					
North lagoon	2 -1285	0.25* - 0.67*	0.50* - 785.50*	0.31 - 2.73	Vitiello & Aïssa, 1985
of Tunis					
Bizerta Bay	14 - 837	0.27 - 1.08	15.25 - 729.25	0.72 - 5.06	Beyrem & Aïssa 2000
South lagoon	1 - 2242	0.24* - 0.85*	8.25* - 1046.50*	0.24 - 2.23	Hermi, 2001
of Tunis					
Ghar El Melh	10 - 210	0.06 - 0.88	1.15 - 183.75	0.56 -3.30	Mahmoudi et al., 2002
lagoon					
Bou Ghrara	154 - 5641	0.19 - 0.81	33.82 - 4569.21	1.22 - 3.76	This work
lagoon					

Bibliographic data on free-living nematode abundance (Individuals 10 cm⁻² (Ind. 10 cm⁻²)), individual weights (µg dry weight (dw)), total biomasse (µg 10 cm⁻² dw) and diversity (H' in bits) in some disturbed areas. * Value given by the Author (s) in wet weight.

of sediments were the most important factors influencing the abundance, biomass and diversity of the nematofauna.

Although a fraction of the organic matter deposited in sediments was labile and hence of potentially high nutritional value for benthic consumers (NEIRA et al., 2001), nematode abundance, mean individual weight, mean total biomass and diversity were negatively and highly correlated with sediment total organic carbon (TOC) amount (p < 0.01). Even though the effects of the organic matter generally, and the TOC in particular. on nematofauna have not been totally clarified, recent studies have indicated that total abundance, diversity and species richness decrease significantly with increasing level of organic enrichment (SCHRATZBERGER & WARWICK, 1998). After the organic loading has eliminated the deeper-living nematodes, the second stage in the faunal succession following the increasing level of the organic load is characterized by the numerical dominance of opportunistic species (PEARSON & ROSENBERG, 1978). These nematode species, better adapted to pronounced unbalanced situations, proliferate and total nematode density can then increase even under polluted conditions.

It is worth mentioning here the much delated Nematode/Copepode ratio of

RAFFAELI-MASON (1981) which was based on this high resistance potentioal of this meiofaunal group.

Similary, negative and highly significant correlations have been noted between abundance, mean individual weight, mean total biomass and Shannon diversity indices (H') of free-living nematode communities and hydrocarbon sediment content (p < 0.01). By increasing the mortality of the more sensitive species (BEYREM & AISSA, 2000; CARMAN et al., 2000), hydrocarbon contamination is responsible for the decrease in nematode densities, total biomass and diversity. This is supported by the fact that the stations richest in hydrocarbons (BG1, BG2, BG3, BG5, BG7 and BG10) are those where we have recorded the lowest values of nematode densities, weights and diversity indices (H'). Besides their direct effect on nematode assemblages, hydrocarbons seem to indirectly influence the nematofauna. So, the accumulation of these pollutants in water entails a reduction its chlorophyll content chlorophyll (BOUCHER et al., 1984), a decrease photosynthetic activity (LACAZE, 1993) and consequently a depletion of dissolved oxygen, a determining factor for quantitative and qualitative data of the

nematofauna (JENSEN, 1987; HEIP, 1995; BEYREM & AISSA, 2000).

Although heavy metal concentrations in the sediment of Bou Ghrara lagoon are lower than the international norms (RAPPORT DGPA, 2000), significant negative correlations have been recorded between these metals (Cu, Pb and Zn) and all indices of nematode communities (p < 0.05). The harmful effect of heavy metals on nematofauna has been demonstrated by many authors and significant negative correlations between the sediment content in these contaminants and nematode densities have been revealed (FICHET & MIRAMAND, 1996; ESSID, 1999). According to BURTON et al. (2001), heavy metal contamination significantly alters the composition of meiofauna assemblages. In Bou Ghrara lagoon, we speculate that the response of nematodes to low heavy metal concentrations is, probably attributed to the fact that these animals have particular abilities to accumulate the Cd, Cu, Pb and Zn present in their biotopes (FICHET et al., 1999).

Among the 11 most dominant species identified, *Terschellingia longicaudata* and *Terschellingia communis* have been found dominant in organically enriched sediments (SCHRATZBERGER & WARWICK, 1998) and both were characterised by low oxygen consumption rates (WARWICK & PRICE, 1998). *Paracomesoma dubium* was recorded abundant in the polluted sediments of the south lagoon (Tunisia) (HERMI, 2001) and *Prochromadorella neapolitana* was found dominant at the polluted sector of the orth lagoon (Tunisia) (VITIELLO & AISSA, 1985).

The free-living nematodes of Bou Ghrara lagoon have showed a decrease in density biomass and diversity in response to pollution stress. These changes in nematofauna community structure highlight the importance of including meiofauna in impact studies of lagoons.

The response of the nematofauna described for the Bou Chrara lagoon corresponds to the

first stage reaction to an incoming pollution stress and maybe refers only to the deep-living nematodes.

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