



Mediterranean Marine Science

Vol 15, No 3 (2014)

Vol 15, No 3 (2014)



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doi: 10.12681/mms.547

To cite this article:

GARCIA-SANZ, S., NAVARRO, P. G., & TUYA, F. (2014). Contrasting recruitment seasonality of sea urchin species in Gran Canaria, Canary Islands (eastern Atlantic). *Mediterranean Marine Science*, *15*(3), 475–481. https://doi.org/10.12681/mms.547

Contrasting recruitment seasonality of sea urchin species in Gran Canaria, Canary Islands (eastern Atlantic)

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Handling Editor: Carlo Nike Bianchi

Received: 19 June 2013; Accepted: 25 December 2013; Published on line: 17 March 2014

Abstract

Although sea-urchins can play an important role affecting the community structure of subtidal bottoms, factors controlling the dynamics of sea-urchin populations are still poorly understood. We assessed the seasonal variation in recruitment of three sea-urchin species (*Diadema africanum*, *Paracentrotus lividus* and *Arbacia lixula*) at Gran Canaria Island (eastern Atlantic) via monthly deployment of artificial collectors throughout an entire annual cycle on each of four adjacent habitat patches (seagrasses, sandy patches, 'urchin-grazed' barrens and macroalgal-dominated beds) within a shallow coastal landscape. *Paracentrotus lividus* and *A. lixula* had exclusively one main recruitment peak in late winter-spring. *Diadema africanum* recruitment was also seasonal, but recruits appeared in late summer-autumn, particularly on 'urchin-grazed' barrens with large abundances of adult conspecifics. In conclusion, this study has demonstrated non-overlapping seasonal recruitment patterns of the less abundant species (*P. lividus* and *A. lixula*) with the most conspicuous species (*D. africanum*) in the study area.

Keywords: recruitment, artificial collectors, coastal habitats, Canary Islands.

Introduction

Sea-urchins are often considered as strong ecological 'interactors', as changes in their abundances can change community composition and structure on inter- and shallow subtidal substrates (Paine, 1992; Sala & Graham, 2002), including overgrazing events with complete denudation of macrophytes (Valentine & Heck, 1999; Tomás et al., 2004). Factors controlling the structure of sea urchin populations are, however, still poorly understood (Norderhaug & Christie, 2009; Uthicke et al., 2009). Several processes, e.g. predation, settlement, recruitment, migration, disease, etc., may play a crucial role in this sense (e.g. Watanabe & Harrold, 1991; Hagen, 1995; McClanahan, 1998; Scheibling et al., 1999; Tomás et al., 2004). Colonization of new habitats by organisms with planktonic larvae, e.g. sea-urchins, is typically divided in four phases: larval development (including dispersal as a planktonic form), testing for habitat suitability, settlement and survival until the organisms might be counted by observers (Underwood, 1979). The number of organisms passing through the fourth phase is defined as recruitment, while the number passing only through the third phase is called settlement (Keough & Downes, 1982). Therefore, recruitment is an interplay of larval and juvenile processes, while settlement involves only larval processes. These early stages are frequently major bottlenecks in the life-history of many benthic invertebrates, including sea-urchins (Gosselin & Qian, 1997; Hunt & Scheibling, 1997; Tomás *et al.*, 2004).

Patterns of distribution and abundance of adult benthic invertebrates can often be determined by recruitment into specific habitats (Underwood & Fairweather, 1989; Roughgarden et al., 1994), although post-recruitment processes, such as competition (Dayton, 1975), predation (Menge, 1976), or physical stress (Garrity, 1984), can modify initial patterns. Variability in recruitment patterns at large spatial scales can result from variations in larval supply, e.g. currents delivering larvae to some locations, but not to others (Minchinton & Scheibling, 1991; Tomás et al., 2004). At smaller scales, active larval choice and preferential recruitment onto the substratum are important in determining the distribution and abundance of benthic organisms, with physical factors such as surface texture, shear stress and patterns of water flow influencing patterns of recruitment (Mullineaux & Butman, 1991; Goldberg & Foster, 2002). For example, when different habitat patches are interspersed within coastal landscapes, the abundance and assemblage structure of new colonizers is affected by the identity of habitat patches (García-Sanz et al., 2012). In the particular case of sea-urchins, processes such as settlement, metamorphosis and recruitment might be induced by the presence of particular algae (Benedetti-Cecchi & Cinelli, 1995; Airoldi, 2000; Bulleri *et al.*, 2002; Privitera *et al.*, 2011) or conspecifics (Cellario & Fenaux, 1990; Gosselin & Jangoux, 1996).

Thorson's (1950) hypothesis points out that temperature causes variation among and within species in larval development and duration. A strong temporal variability is a general feature in sea urchin settlement and recruitment in both temperate (e.g. Ebert et al., 1994; Balch & Scheibling, 2000; Hereu et al., 2004; Hereu et al., 2012) and subtropical and tropical seas (e.g. Keesing et al., 1993; Hernández et al., 2006), including seasonal patterns in recruitment every year following specific thermal patterns. According to Hoegh-Guldberg & Pearse (1995), sea-water temperature, rather than food or any other mechanism, best explains development of marine invertebrates, sea-urchins in particular. On the one hand, temperature enhances echinoderm settlement by accelerating larval development and reducing the period during which larvae are exposed to planktonic predators (Hart & Scheibling, 1988; Hoegh-Guldberg & Pearse, 1995). On the other hand, cold nutrient-rich waters may trigger phytoplankton blooms that are thought to be synchronized with sea-urchin spawning, thereby improving their reproductive success (Starr et al., 1990) and subsequent settlement and recruitment of certain species (Ebert, 1983).

In the Canary Islands (eastern Atlantic), three sea urchin species co-occur in the rocky subtidal from 0 to 30 m depth: *Diadema africanum*, *Paracentrotus lividus* and *Arbacia lixula*. Both *P. lividus* and *A. lixula* are well-adapted to the turbulent waters of the first few meters of the subtidal, while *D. africanum* commonly increases in abundance with depth (Tuya *et al.*, 2007). The locally abundant *D. africanum* occurs

in the Eastern Atlantic islands (Rodríguez et al., 2013), from Madeira to the Guinean Gulf, including the Salvage Islands, Canary Islands, Cape Verde Islands (Hernández et al., 2008), and São Tome Island (Lessios et al., 2001). It has also been recorded in continental areas of Ghana (John et al., 1977, 1992) and in Ngor Island, Senegal (P. Wirtz, pers. com). On the other hand, P. lividus has an Atlanto-Mediterranean distribution (Boudouresque & Verlaque, 2001), including panmixia within the Atlantic and Mediterranean basins (Durán et al., 2004; Calderón et al., 2008). In the Atlantic, this sea urchin is distributed from Ireland to the Canary Islands (southern range edge), therefore showing an affinity for temperate and warmtemperate waters. Finally, A. lixula distribution includes the African Atlantic coast from Morocco to Angola, the east Atlantic archipelagos of Cape Verde, the Canaries, Madeira and the Azores, and the whole Mediterranean basin, excluding the Black Sea (Wangensteen et al., 2012).

In this study, we aimed to assess whether recruitment patterns of three sea urchin species (*D. africanum*, *P. lividus* and *A. lixula*) differed from season to season through an annual cycle between adjacent habitats within a coastal landscape.

Materials and Methods

Study area

This study was carried out at two localities off the east coast of Gran Canaria (Canary Islands, Spain): Playa del Cabrón ($27^{\circ}52'14.43''$ N, $15^{\circ}23'00.31''$ W) and Risco Verde ($27^{\circ}51'25.94''$ N, $15^{\circ}23'10.26''$ W) (Fig. 1). These sites encompass a mosaic of subtidal, neighboring habitats, including seagrass meadows constituted by *Cy*-



Fig. 1: Location of the study area in the eastern Atlantic Ocean, Playa del Cabrón and Risco Verde off Gran Canaria Island.

modocea nodosa and sandy patches on soft substrata, as well as 'urchin-grazed' barrens mainly generated by large densities of the sea urchin *D. africanum* and macroalgaldominated beds (primarily constituted by the frondose genera *Cystoseira* and *Sargassum*, as well as *Dyctiota* spp., *Padina pavonica* and *Lobophora variegata*) on rocky substrates. All four habitats are found interspersed as a matrix of mosaic patches at 5-10 meters depth with similar orientation and exposure to waves.

Collector deployment

Recruitment of D. africanum, P. lividus and A. lixula was assessed through artificial collectors (a cushion-shaped leaf-like unit), which has been previously shown to attract postlarvae and newly metamorphosed juveniles of seaurchins in the study area (García-Sanz et al., 2012). This collector consists of a plastic mesh frame (50 cm², 2 cm diameter of mesh size) to which artificial leaves (green plastic raffia, 35 cm long and 10 mm wide) are attached every ca. 4 cm, and folded as a cushion. Raffia is positively buoyant under water and consequently tends to float upright. A total of 75 cm² of concealment gardening mesh (≤ 1 mm diameter) was included inside, creating small holes and shelters (for an image of the collector, see García-Sanz et al., 2012). Four artificial substrates (ASs) were deployed, on a monthly basis through one year (from February 2009 to February 2010), on each of the four habitats, at each locality. Adjacent ASs within each type of habitat were ~ 5 m apart; ASs from any two adjacent habitats were 100s of m apart. ASs were fixed through cable ties on hard substrates and by iron rod (ca. 50 cm long) on soft substrates. ASs were in contact with the seabed. Collectors were retrieved, in all cases, after a period of 4 weeks. Each AS was removed by scuba-divers, by carefully enclosing each unit within a cloth bag. The bags were carried to the laboratory, where each collector was cleaned with fresh water, and all sea-urchins retained by a 0.5 mm mesh sieve subsequently identified and counted.

Data analysis

A large number of ASs was lost and/or stolen during the study. As a result, locality (a random source of variability) was 'sacrificed' to maintain a balanced design by pooling AS from each month and habitat from both localities; the 4 replicates of each AS from each month and habitat are considered as independent replicates.

Differences in the total abundance of *D. africanum*, *P. lividus* and *A. lixula* recruits between seasons (winter, spring, summer and autumn), months within seasons and habitats were separately tested through a 3-way, permutation-based (via 4999 random permutations), ANOVA based on Euclidean distances calculated from square-root transformed data. The model included the fixed factors: 'Season' and 'Habitat', while 'Month' was a random factor nested within 'Season'. Pairwise comparisons (through 4999 random permutations) were used

to resolve differences among levels of factors, whenever appropriate, particularly when a significant 'Season x Habitat' was detected.

Results

A total of 291 sea urchin recruits (larvae and juveniles) were collected during the study (226 *P. lividus*, 38 *D. africanum* and 27 *A. lixula*). Although no significant differences were found for season (Table 1), *P. lividus* had a main recruitment peak in late winter-spring, while only a few colonizers were observed during late summer and early autumn (Fig. 2a). During the recruitment peak, significant differences in the abundance of *P. lividus* recruits were found between habitats ('Habitat', F= 3.028, P= 0.001, Table 1); soft-substrate habitats (sandy patches and seagrasses) showed larger abundances than hard-substrate habitats ('urchin-grazed' barrens and macroalgal beds) (Fig. 2a, pairwise tests, Table 1), but not in the other seasons.

Recruitment of *D. africanum* was also seasonal, with recruits, however, appearing in autumn-winter (Fig. 2; Table 1, 'Season', F=3.314, P=0.016), showing differences between months in season (Fig. 2; Table 1, 'Month (Season)', F=2.311, P=0.018). The abundance of *D. africanum* recruits was larger in 'urchin-grazed' barrens and sandy patches than in the other 2 habitats (Fig. 2.b, 'Habitat' F=10.112, P=0.001, pairwise comparisons, Table 1). Seasonal differences in the abundance of *D. africanum* recruits were inconsistent between habitats ('Season x Habitat', F=4.681, P=0.001, Table 1).

Arbacia lixula showed a main recruitment peak in late winter-spring (Fig. 2c) that resulted in marginally significant differences between seasons (Table 1, 'Season', F=3.057, P=0.063); that may be due to significant differences observed for month in season (Table 1, 'Month (Season)', F=3.661, P=0.001). This seasonal pattern was irrespective of the type of habitat ('Season x Habitat', F=0.581, P=0.812, Table 1).

Discussion

Our study has shown that *D. africanum* has a clear unimodal recruitment pattern, with a main peak occurring in autumn (September, October and November); this has been previously reported for Tenerife Island (Canary Islands, Hernández *et al.*, 2010), which typically occurs 2 months after the maximum spawning period in the study area, i.e. April–June (Hernández *et al.*, 2006). The recruitment of *D. africanum* is, therefore, negatively related to food availability in the water column, i.e. phytoplankton availability in the 2 months preceding settlement, which indicates that *D. africanum* larvae are able to reach competence when limited food is available (Hernández *et al.*, 2010). Indeed, *Diadema* species are able to survive and develop in low-food environments (McAlister, 2008), and our results reinforce the notion that *D. africanum* is

Table 1. Results of 3-way, permutation-based, ANOVA testing for differences in the abundance of *Diadema africanum*, *Paracentrotus lividus* and *Arbacia lixula* recruits between seasons (Winter = W, Spring =Sp, Summer = S and Autumn = A), months within seasons and habitats (Barrens = B, Macroalgal Beds = MB, Sandy Patches = SP and Seagrasses = S). Pairwise comparisons for significant 'Habitat', 'Season' and 'Season x Habitat' interactions are included.

	df	MS	F	Р		Pairwise comparisons
Paracentrotus lividus						
Season	3	8.286	1.575	0.261		
Habitat	3	2.505	3.028	0.044	Sp > S = B = MB	
Month (Season)	8	5.258	8.161	0.001		
Season x Habitat	9	0.793	0.958	0.505		
Month (Season) x Habitat	24	0.827	1.284	0.159		
Residual	144	0.644				
Diadema africanum						
Season	3	1.284	3.314	0.016	A > W > S = Sp	Season x Habitat
Habitat	3	0.491	10.112	0.001	B > SP > M > S	W: $B > SP > MB = S$
Month (Season)	8	0.387	2.311	0.018		Sp: $B = SP = MB = S$
Season x Habitat	9	0.227	4.681	0.001		S: B = SP = MB = S
Month (Season) x Habitat	24	0.048	0.29	0.999		A: $B > SP > MB > S$
Residual	144	0.167				
Arbacia lixula						
Season	3	1.004	3.057	0.063		
Habitat	3	0.188	0.967	0.433	S = SP > B > M	
Month (Season)	8	0.328	3.661	0.001		
Season x Habitat	9	0.113	0.581	0.812		
Month (Season) x Habitat	24	0.195	2.175	0.004		
Residual	144	0.089				

well-adapted to the oligotrophic waters around the Canaries (Hernández et al., 2006).

This study has also shown one main recruitment peak in late winter-early spring (February, March and April) for both P. lividus and A. lixula in Gran Canaria Island. This recruitment peak coincides with the lower temperatures observed during the year (< 20°C) and the period of maximum primary production in the study area (Hernández et al., 2006). A priori, this suggests that pelagic larvae of P. lividus and A. lixula require cold, nutrient-rich, waters for their development, metamorphosis and subsequent recruitment into the bottom. This outcome coincides with the main recruitment peak observed for P. lividus during spring and early summer in Mediterranean areas with similar temperature regimes (Azzolina & Willsie, 1987; Lozano et al., 1995; Sala & Zabala, 1996; Hereu et al., 2004; Tomás et al., 2004). We also observed a secondary recruitment peak in early autumn (September-October), which coincides with similar observations from the Mediterranean Sea (Verlague, 1984; Guettaf et al., 2000). At Tenerife Island (Canary Islands), two peaks of gonadal maturation have been observed for this echinoid during winter (December) and early summer (June-July), respectively (Girard et al., 2006); however, the arrival of new recruits associated with the summer peak has not been recorded (Girard et al., 2006). At Gran Canaria Island, a few recruits were observed corresponding with the second, early summer peak. This could

be explained since offshore waters around Gran Canaria are typically colder and richer in primary productivity than in Tenerife, as a result of the oceanographic east-towest gradient crossing the Canarian Archipelago (Barton *et al.*, 1998). Most likely, this guarantees competence and survivorship of *P. lividus* larvae in the water column during summer. In the particular case of *A. lixula*, our results (optimal recruitment at 18°C) clearly contrast with the temperature (22°C) identified in the Mediterranean at which larvae optimize survivorship through metamorphosis (Privitera et al. 2011). This discrepancy points towards potential differences at which recruits reach optimum competence across the species' distribution range.

It is widely known that processes such as settlement, larval metamorphosis and recruitment of sea-urchins are often induced and/or facilitated by the occurrence of adult conspecifics (Cellario & Fenaux, 1990; Gosselin & Jangoux, 1996). Such a pattern has been observed here for *D. africanum* recruits, which reached larger abundances in 'urchin-grazed' barrens than in the other habitat patches. In contrast, the abundance of both *P. lividus* and *A. lixula* recruits was not higher in those habitat patches where adults have been typically found in the study area, i.e. rocky substrates at shallow water dominated by macroalgal beds (Tuya *et al.*, 2007). Most likely, *Diadema* recruits have a tendency to look for protection under the long-spined 'canopy' provided by adult conspecifics (Atkinson *et al.*, 1973), and so increase their



Fig. 2: Abundance of (a) *Paracentrotus lividus*, (b) *Diadema africanum* and (c) *Arbacia lixula* recruits in artificial collectors per month and habitat. The mean sea water temperature is also included. Error bars are + SE of means.

survivorship by releasing predation, which notoriously can affect Diadema recruits in the study area (Clemente et al., 2007). Although such a 'canopy' effect has also been described for some short-spined sea-urchins (Tegner & Dayton, 1981), this mechanism seems irrelevant for both P. lividus and A. lixula, i.e. two short-spined sea urchin species. Across the Canary Islands, these two sea-urchins have overall adult abundances considerably lower than D. africanum (Tuya et al., 2007), which is the species that majorly generates 'grazed urchin' barrens (Tuya et al., 2004; Clemente et al., 2007). The low abundances of both P. lividus and A. lixula do not seem to promote a 'canopy' effect on small-sized individuals, as a way to provide shelter to new recruits. Hence, this might help to explain the lack of recruits of both P. lividus and A. lixula in collectors deployed on rocky substrates.

Contrary to our expectations, *P. lividus* recruits had larger abundances on sandy patches relative to the other habitats. This result may be an artifact caused by the presence of collectors on bottoms that lack provision of any shelter, i.e. an 'oasis' effect: attraction of new individuals to the only available substrate (García-Sanz *et al.*, 2012). Indeed, individuals of *P. lividus* can easily move among adjacent habitat patches (Ceccherelli *et al.*, 2009), which could guarantee further arrival of individuals to favorable rocky substrates covered by macroalgae (Tuya *et al.*, 2007). In conclusion, this study has demonstrated non-overlapping seasonal recruitment patterns of the less abundant species (*P. lividus* and *A. lixula*) with the most conspicuous species (*D. africanum*) of the study area.

Acknowledgements

We thank J. Suárez for providing help in the laboratory and "Puertos del Estado", who kindly provided the sea water temperature data sets. SGS was financed through a postgraduate fellow grant provided by the Cabildo de Gran Canaria. Ricardo Haroun provided logistical support during the lab work.

References

- Airoldi, L., 2000. Responses of algae with different life histories to temporal and spatial variability of disturbance in subtidal reefs. *Marine Ecology Progress Series*, 195, 81-92.
- Atkinson, C., Hopley, S., Mendelsohn, L., Yacowitz, S., 1973. Food studies on *Diadema antillarum* on a patch reef. p. 65-80. In: *Studies on the activity and food of the echinoid Diadema antillarum Philippi on a West Indian Patch Reef, St Croix,* US Virgin Islands. Ogden, J.D., Abbott, D.P., Abbott, I. (Eds). West Indies Lab., St Croix.
- Azzolina, J.F., Willsie, A., 1987. Abundance des juveniles de *Paracentrotus lividus* au sein de l'herbier a *Posidonia oceanica*.
 p. 159-167. In: *Colloque international sur Paracentrotus lividus et les oursins comestibles*. Boudouresque, C.F. (Ed.). GIS Posidonie, Marseille.
- Balch, T., Scheibling R.E, 2000. Temporal and spatial variability in settlement and recruitment of echinoderms in kelp beds and barrens in Nova Scotia. *Marine Ecology Progress Series*, 205, 139-154.
- Barton, E.D., Arístegui, J., Tett, P., Cantón, M., García-Braun, J. et al., 1998. The transition zone of the Canary Current upwelling region. *Progress in Oceanography*, 41, 455-504.
- Benedetti-Cecchi, L., Cinelli, F., 1995. Habitat heterogeneity, sea urchin grazing and the distribution of algae in littoral rock pools on the west coast of Italy (western Mediterranean). *Marine Ecology Progress Series*, 126, 203-212.
- Boudouresque, C.F., Verlaque, M., 2001. Ecology of Paracentrotus lividus. p. 177-216. In: Edible sea-urchins: biology and ecology. Lawrence, J.M. (Ed.). Elsevier Science, Amsterdam.
- Bulleri, F., Bertoci, I., Micheli, F., 2002. Interplay of encrusting coralline algae and sea-urchins in maintaining alternative habitats. *Marine Ecology Progress Series*, 243, 101-109.Calderón, I., Giribet, G., Turón, X., 2008. Two markers and one history: phylogeography of the edible common sea urchin *Paracentrotus lividus* in the Lusitanian region. *Marine Biology*, 154, 137-151.
- Ceccherelli, G., Pais, A., Pinna, S., Serra, S., Sechi, N., 2009. On the movement of the sea urchin *Paracentrotus lividus* towards *Posidonia oceanica* seagrass patches. *Journal of Shellfish Research*, 28, 397-403.
- Cellario, C., Fenaux, L., 1990. *Paracentrotus lividus* (Lamarck) in culture (larval and benthic phases): parameters of growth observed during two years following metamorphosis. *Aquaculture*, 84, 173-188.
- Clemente, S., Hernández, J.C., Toledo, K., Brito, A., 2007. Predation upon *Diadema* aff. *antillarum* in barren grounds in the Canary Islands. *Scientia Marina*, 71, 745-754.
- Dayton, P.K., 1975. Competition, disturbance and community organization: the provision and subsequent utilization of space in a rocky intertidal community. *Ecological Monographs*, 41, 351-389.
- Durán, S., Palacín, C., Becerro, M.A., Turón, X., Giribet, G., 2004. Genetic diversity and population structure of the commercially harvested sea urchin *Paracentrotus lividus* (Echinodermata, Echinoidea). *Molecular Ecology*, 13, 3317-3328.
- Ebert, T.A., 1983. Recruitment in echinoderms. p. 169-203. In: *Echinoderm studies*. Lawrence, J.M., Jangoux, M. (Eds). AA Balkema, Rotterdam.
- Ebert, T.A., Schroeter, S.C., Dixon, J.D., Kalvass, P., 1994. Settlement patterns of red and purple sea-urchins (*Strongylocen*-

trotus franciscanus and *S. purpuratus*) in California, USA. *Marine Ecology Progress Series*, 111, 41-52.

- García-Sanz, S., Tuya, F., Navarro, P.G., Angulo-Preckler, C., Haroun, R.J., 2012. Post larval, short-term, colonization patterns: The effect of substratum complexity across subtidal, adjacent, habitats. *Estuarine Coastal and Shelf Science*, 112, 183-191.
- Garrity, S.D., 1984. Some adaptations of gastropods to physical stress on a tropical rocky shore. *Ecology*, 65, 559-574.
- Girard, D., Hernández, J.C., Toledo, K., Clemente, S., Brito, A., 2006. Aproximación a la biología reproductiva del equinodermo *Paracentrotus lividus* (Lamarck, 1816) en el litoral de Tenerife. In: XIV Simposio Ibérico de Estudios de Biología Marina, 12th-15th September 2006. Barcelona, España.
- Goldberg, N.A., Foster, M.S., 2002. Settlement and postsettlement processes limit the abundance of the geniculate coralline alga Calliarthron on subtidal walls. *Journal of Experimental Marine Biology and Ecology*, 278, 31-45.
- Gosselin, P., Jangoux, M., 1996. Induction of metamorphosis in *Paracentrotus lividus* larvae (Echinodermata, Echinoidea). *Oceanologica Acta*, 19, 293-296.
- Gosselin, L.A., Qian, P.Y., 1997. Juvenile mortality in benthic marine invertebrates. *Marine Ecology Progress Series*, 146, 265-282.
- Guettaf, M., San Martin, G.A., Francour, P., 2000. Interpopulation variability of the reproductive cycle of *Paracentrotus lividus* (Echinodermata: Echinoidea) in the south-western Mediterranean. *Journal of the Marine Biological Association of the* U. K., 80, 899-907.
- Hagen, N.T., 1995. Recurrent destructive grazing of successional immature kelp forest by green sea-urchins in Vestfjorden, Northern Norway. *Marine Ecology Progress Series*, 123, 95-106.
- Hart, M.W., Scheibling, R.E., 1988. Heat waves, baby booms, and the destruction of kelp beds by sea-urchins. *Marine Biology*, 99, 167-176.
- Hereu, B., Zabala, M., Linares, C., Sala, E., 2004. Temporal and spatial variability in settlement of the sea urchin *Paracentrotus lividus* in the NW Mediterranean. *Marine Biology*, 144, 1011-1018.
- Hereu, B., Linares, C., Sala, E., Garrabou, J., Garcia-Rubies, A. *et al.*, 2012. Multiple processes regulate long-term population dynamics of sea-urchins on Mediterranean rocky reefs. PLoS ONE 7(5): e36901. doi:10.1371/journal.pone.0036901.
- Hernández, J.C., Brito, A., Cubero, E., Girard, D., González-Lorenzo, G. *et al.*, 2006. Temporal patterns of larval settlement of *Diadema antillarum* (Echinodermata: Echinoidea) in the Canary Islands using an experimental larval collector. *Bulletin of Marine Science*, 78, 271-279.
- Hernández, J.C., Clemente, S., Sangil, C., Brito, A., 2008. Actual status of the sea urchin *Diadema* aff.*antillarum* populations and macroalgal cover in the Marine Protected Areas comparing to a highly fished area (Canary Islands-Eastern Atlantic Ocean). *Aquatic Conservation: Marine and Freshwater Research*, 66, 259-270.
- Hernández, J.C., Clemente, S., Girard, D., Perez-Ruzafa, A., Brito, A., 2010. Effect of temperature on settlement and postsettlement survival in a barren-forming sea urchin. *Marine Ecol*ogy Progress Series, 413, 69-80.
- Hoegh-Guldberg, O., Pearse, J.S., 1995. Temperature, food availability, and development of marine invertebrate larvae. *American Zoologist*, 35, 415-425.
- Hunt, H.L., Scheibling, R.E., 1997. Role of early post-settlement

mortality in recruitment of benthic marine invertebrates. *Ma*rine Ecology Progress Series, 155, 269-301.

- John, D.M., Lieberman, D., Lieberman, M., 1977. A Quantitative study of the structure and dynamics of benthic subtidal algal vegetation in Ghana (Tropical west Africa). *Journal of Ecol*ogy, 65, 497-521.
- John, D.M., Price, J.H., Lawson, G.W., 1992. Tropical east Atlantic and Islands: plant-animal interactions on tropical shores free of biotic reef. p. 87–99. In: *Animal Interactions in the Marine Benthos*. Jonh, D.M., Hawkins, S.J., Price, J.H. (Eds). Oxford Clarendon Press Oxford.
- Keesing, J.K., Cartwright, C.M., Hall, K.C., 1993. Measuring settlement intensity of echinoderms on coral reefs. *Marine Biol*ogy, 117, 399-407.
- Keough, M.J., Downes, B.J., 1982. Recruitment of marine invertebrates: the role of active choices and early mortality. *Oecologia*, 54, 348-352.
- Lessios, H.A, Kessing, B.D., Pearse, J.S., 2001. Population structure and speciation in tropical seas: global phylogeography of the sea urchin Diadema. *Evolution*, 55, 955-975.
- Lozano, J., Galera, J., López, S., Turón, X., Palacín, C. et al., 1995. Biological cycles and recruitment of *Paracentrotus lividus* (Echinodermata: Echinoidea) in two contrasting habitats. *Marine Ecology Progress Series*, 122, 179-191.
- McAlister, J.S., 2008. Evolutionary response to environmental heterogeneity in central American echinoderm larvae: plastic versus constant phenotypes. *Evolution*, 62, 1358-1372.
- McClanahan, T.R., 1998. Predation and the distribution and abundance of tropical sea urchin populations. *Journal of Experimental Marine Biology and Ecology*, 221, 231-255.
- Menge, B.A., 1976. Organization of the New England rocky intertidal community: role of predation, competition, and environmental heterogeneity. *Ecological Monographs*, 46, 355-393.
- Minchinton, T.E., Scheibling, R.E., 1991. The influence of larval supply and settlement on the population structure of barnacles. *Ecology*, 72, 1867-1879.
- Mullineaux, L.S., Butman, C.A., 1991. Initial contact, exploration and attachment of barnacle (*Balanus amphitrite*) cyprids settling in flow. *Marine Biology*, 110, 93-103.
- Norderhaug, K.M., Christie, H.C., 2009. Sea urchin grazing and kelp re-vegetation in the NE Atlantic. *Marine Biology Research*, 5, 515-528.
- Paine, R.T., 1992. Food-web analysis through field measurements of per capita interaction strength. *Nature*, 355, 73-75.
- Privitera, D., Noli, M., Falugi, C., Chiantore, M., 2011. Benthic assemblages and temperature effects on *Paracentrotus lividus* and *Arbacia lixula* larvae and settlement. *Journal of Experimental Marine Biology and Ecology*, 407, 6-11.
- Rodríguez, A., Hernández, J.C., Clemente, S., Coppard, S.E., 2013. A new species of *Diadema* (Echinodermata: Echinoidea: Diadematidae) from the Eastern Atlantic Ocean and a neotype designation of *Diadema antillarum* (Philippi, 1845). *Zootaxa*, 3636, 144-170.
- Roughgarden, J., Pennington, T., Alexander, S., 1994. Dynamics of the rocky intertidal zone with remarks on the generalization in ecology. *Philosophical Transactions of the Royal Society of London B*, 343, 79-85.

- Sala, E., Zabala, M., 1996. Fish predation and the structure of the sea urchin *Paracentrotus lividus* populations in the NW Mediterranean. *Marie Ecology Progress Series*, 140, 71-81.
- Sala, E., Graham, M.H., 2002. Community-wide distribution of predator-prey interaction strength in kelp forests. *Proceedings of the National Academy of Sciences USA*, 99, 3678-3683.
- Scheibling, R.E., Hennigar, A.W., Balch, T., 1999. Destructive grazing, epiphytism, and disease: the dynamics of sea urchinkelp interactions in Nova Scotia. *Canadian Journal of Fisheries and Aquatic Sciences*, 56, 2300-2314.
- Starr, M., Himmelman, J.H., Therriault, J.C., 1990. Direct coupling of marine invertebrate spawning with phytoplankton blooms. *Science*, 247, 1071-1074.
- Tegner, M.J., Dayton, P.K., 1981. Population structure, recruitment and mortality of two sea-urchins *Strongylocentrotus franciscanus* and *S. purpuratus* in a kelp forest. *Marie Ecology Progress Series*, 5, 255-268.
- Thorson, G., 1950. Reproduction and larval ecology of marine bottom invertebrates. *Biological Reviews of the Cambridge Philosophical Society*, 25, 1-45.
- Tomás, F., Romero, J., Turón, X., 2004. Settlement and recruitment of the sea urchin *Paracentrotus lividus* in two contrasting habitats in the Mediterranean. *Marie Ecology Progress Series*, 282, 173-184.
- Tuya, F., Boyra, A., Sánchez-Jerez, P., Barberá, C., Haroun, R.J., 2004. Can one species determine the structure of the benthic community on a temperate rocky reef? The case of the longspined sea-urchin *Diadema antillarum* (Echinodermata: Echinoidea) in the eastern Atlantic. *Hydrobiologia*, 519, 211-214.
- Tuya, F., Cisneros-Aguirre, J., Ortega-Borges, L., Haroun, R.J., 2007. Bathymetric segregation of sea-urchins on reefs of the Canarian Archipelago: Role of flow-induced forces. *Estuarine, Coastal and Shelf Science*, 73, 481-488.
- Underwood, A.J., 1979. The ecology of intertidal gastropods. *Advances in Marine Biology*, 16, 111-210.
- Underwood, A.J., Fairweather, P.J., 1989. Supply-side ecology and benthic marine assemblages. *Trends in Ecology and Evolution*, 4, 16-20.
- Uthicke, S., Schaffelke, B., Byrne, M., 2009. A boom–bust phylum? Ecological and evolutionary consequences of density variations in echinoderms. *Ecological Monographs*, 79, 3-24.
- Valentine, J.F., Heck Jr. K.L., 1999. Seagrass herbivory: evidence for the continued grazing of marine grasses. *Marie Ecology Progress Series*, 176, 291-302.
- Verlaque, M., 1984. Biologie des juvéniles de l'oursin herbivore Paracentrotus lividus (lamarck): séléctivité dur broutage et impact de l'espèce sur les communautés algales de substrat rocheux en Corse (Méditerranée, France). Botanica Marina, 27, 401-424.
- Wangensteen, O.S., Turon, X., Pérez-Portela, R., Palacín, C., 2012. Natural or naturalized? phylogeography suggests that the abundant sea urchin *Arbacia lixula* is a recent colonizer of the Mediterranean. *PLoS ONE* 7 (9), e45067.
- Watanabe, J., Harrold, C., 1991. Destructive grazing by sea-urchins *Strongylocentrotus* spp. in a central California kelp forest: potential roles of recruitment, depth, and predation. *Marine Ecology Progress Series*, 71, 125-141.