

The deep vault: a temporary refuge for temperate gorgonian forests facing marine heat waves

Lorenzo BRAMANTI¹, Elisabetta MANEA¹, Bruna GIORDANO^{1,2}, Tristan ESTAQUE³, Olivier BIANCHIMANI³, Justine RICHAUME³, Bastien MÉRIGOT⁴, Quentin SCHULL⁴, Stephane SARTORETTO⁵, Joaquim GARRABOU⁶ and Katell GUIZIEN¹

¹ Laboratoire d'Ecogéochimie des Environnements Benthiques, LECOB, Observatoire Océanologique de Banyuls sur Mer Centre National de la Recherche Scientifique (CNRS)-Sorbonne Université, Banyuls sur Mer, France

² Department of Life and Environmental Sciences, University of Cagliari, Cagliari, Italy

³ Septentrion Environnement, Campus Nature Provence, 89 Traverse Parangon, Marseille, France

⁴ MARBEC, Univ. Montpellier, IFREMER, IRD, CNRS, Sète, France

⁵ Ifremer, LITTORAL, F-83500 La Seyne-sur-Mer, France

⁶ Instituto de Ciencias del mar ICM-CSIC, Barcelona

Corresponding author: Elisabetta MANEA; elisabetta.manea@obs-banyuls.fr

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Abstract

Climate change poses a significant threat to coastal areas, marked by the increasing intensity and frequency of marine heat waves observed in various ecosystems around the world. Over the last 25 years, a vast number of Mediterranean populations of the red gorgonian *Paramuricea clavata* have been impacted by marine heatwaves. The last mass mortality occurred during the summer of 2022 in the Western Mediterranean Sea, affecting mostly shallow populations (down to 30 m depth). Here we provide an assessment of the health status of mesophotic *P. clavata* populations down to 90 m depth to investigate a depth refuge hypothesis. Results show that the impact of marine heat waves decreases with depth, with a significant drop in mortality below 40 m depth. These observations support the hypothesis of a depth refuge from marine heat waves that may allow, at least temporarily, the maintenance of *P. clavata* in the Western Mediterranean Sea. The present study strongly advocates for further investigations and monitoring of the mesophotic zone to chart potential areas that could serve as deep refuge for gorgonians.

Keywords: Mass mortality events; sea-fan; Mediterranean Sea; mesophotic; deep refuge; extreme events; temperate corals; habitat-forming species; climate change; octocorals.

Introduction

Since 1990, marine heatwaves (MHWs) have increased in frequency, extent, and intensity worldwide, causing major changes in marine ecosystems (Dayan *et al.*, 2023; Smith *et al.*, 2023). Usually, when high and extreme temperatures are recorded, they are associated with intense mass mortalities of marine benthic organisms. A case in point is the red gorgonian *Paramuricea clavata* (Risso, 1827), one of the most affected species in the Mediterranean Sea during the summer period (July-August) (e.g., Rivetti *et al.*, 2014; Garrabou *et al.*, 2022; Estaque *et al.*, 2023). This gorgonian is a keystone engineering species, offering a forest-like habitat with a high three-dimensional complexity, especially when its populations reach high densities of large colonies (>20 cm high) (Rossi *et al.*, 2017). Numerous vertebrates and invertebrates find shelter and food under their canopies (Ponti *et al.*, 2018). Therefore, the fate of these forests

is of great concern, as their degradation leads to strong impacts on the whole ecosystem with consequent loss of biodiversity and ecosystem services (Ponti *et al.*, 2014; Paoli *et al.*, 2017; Garrabou *et al.*, 2021; Gómez-Gras *et al.*, 2021; Rossi *et al.*, 2022).

The first mass mortality events of *P. clavata* were reported in the 1980's and generally attributed to pollution, sometimes related to heavy rainfalls (see for instance Arnoux *et al.*, 1992; Bavestrello *et al.*, 1994). These relatively localized events affected both shallow and deep populations (down to 80 m). Since 1999, repeated mass mortality events have been recorded (2003, 2006, 2018, and 2022), mainly in shallow populations of the Western Mediterranean, affecting the coasts from Spain through France up to Italy, as well as the Adriatic Sea. The potential cause behind these drastic events has been attributed to temperature anomalies associated with MHWs (Cerrano *et al.*, 2000; Perez *et al.*, 2000; Garrabou *et al.*, 2009, 2022; Grenier *et al.*, 2023). Since then, researchers have been raising the

alarm as current demographic trajectories of population decline could ultimately lead to the collapse of *P. clavata* populations, assuming that their resilience potential will decrease within the ongoing warming context (Garrahou *et al.*, 2021). Nonetheless, some cases of population recovery have been recorded after two well documented mass mortality events in 1999 and 2003 (Cupido *et al.*, 2009; Santangelo *et al.*, 2015), with genetic studies highlighting the putative role of connectivity between shallow and deep populations as a driver of recovery success (Pilczynska *et al.*, 2016; Padrón *et al.*, 2018a).

In the summer of 2022, the Mediterranean was hit by one of the strongest MHWs ever recorded, affecting its western basin (Guinaldo *et al.*, 2023), followed by an extensive red gorgonian mortality that began in August 2022 (Estaque *et al.*, 2023). According to published observations, mortality rates were extreme down to 30 m (Estaque *et al.*, 2023). Also other species of gorgonians, corals, algae, bryozoans, and sponges were severely affected, with bath sponges reported to be on the brink of local extirpation along the entire French Mediterranean coast (Grenier *et al.*, 2023). Unlike *P. clavata*, whose habitat extends to depths of 200 m (Carpine & Grasshoff, 1975; Weinberg, 1976), the affected bath sponges on the French coast typically dwell in shallow water (10-20 m). The extirpation of this shallow water sponges populations

matches *P. clavata* mortality events which have been reported almost exclusively for shallow populations (15-25 m depth, rarely >30). The recent emergence of citizen science initiatives has led to increased early detections of these phenomena available and generated a wealth of data that were very difficult to obtain 20 years ago (Kelly *et al.*, 2020; Vicioso *et al.*, 2022). Nonetheless, these initiatives usually involve recreational SCUBA divers (Turicchia *et al.*, 2021; Figuerola-Ferrando *et al.*, 2022), who rarely venture below 30 m. Hence, while monitoring efforts at depths below 30 m remain limited, the few existing observations suggest only a minor or no impact on *P. clavata* below 40 m (Arnoux *et al.*, 1992; Linares *et al.*, 2005; Pérez-Portela *et al.*, 2016; Estaque *et al.*, 2023). Since the severity of these events seems to be closely linked to the duration of high temperatures exposure, and thus, to the depth of the thermocline (Pérez *et al.*, 2006 and author's personal observations), more monitoring efforts should focus on deeper *P. clavata* populations to improve our understanding of the causal relationships between MHWs and species die-off, especially with regard to the potential of *P. clavata* to resist to MHWs effects at greatest depths. Indeed, the wide bathymetric distribution of *P. clavata* offers a glimmer of hope against the rather bleak backdrop of seemingly inevitable and dramatic future temperature increase.

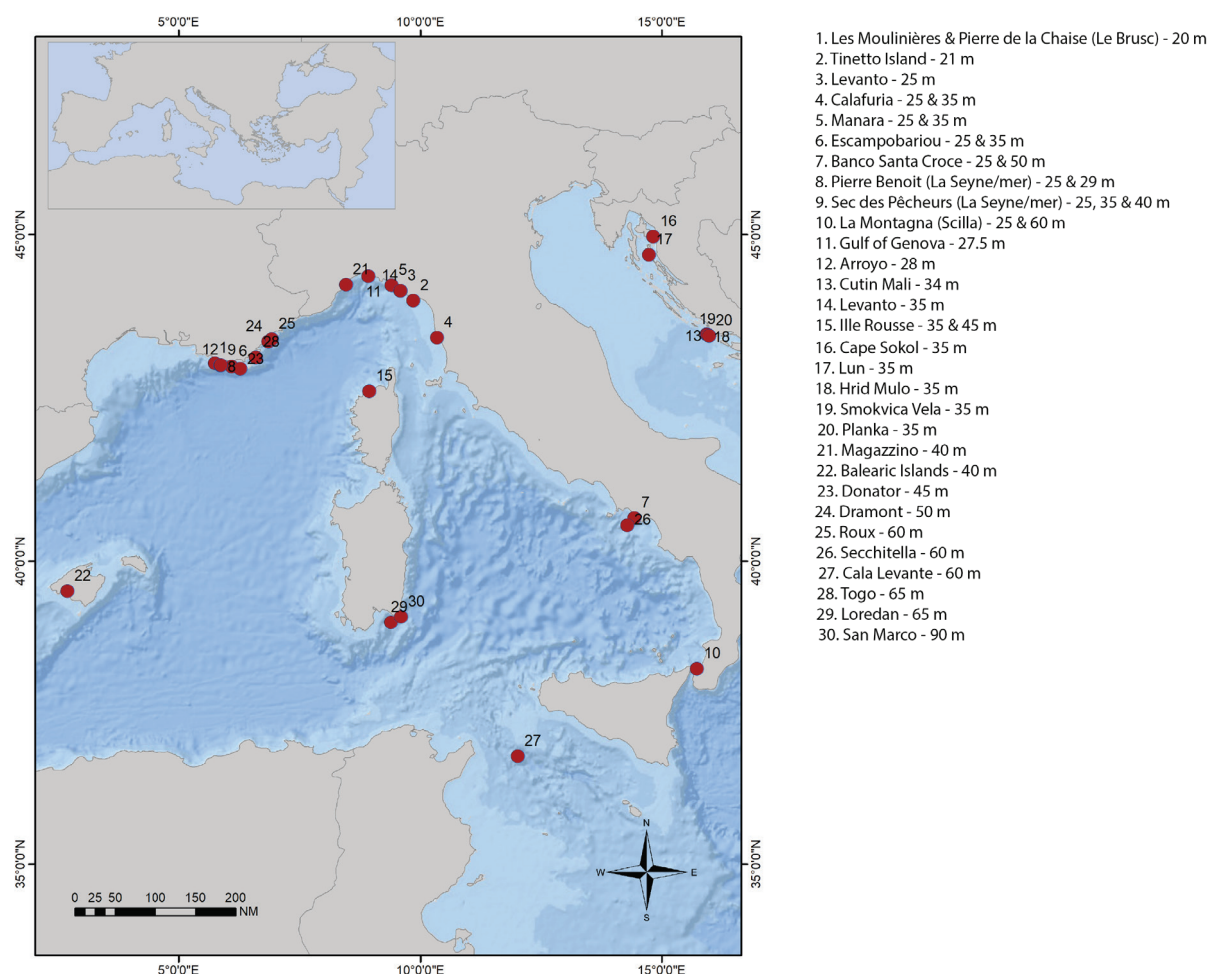


Fig. 1: Map of the studied sites. The names of the sites and the depth (m) at which data on the red gorgonian *Paramuricea clavata* health status were recorded are reported in the legend map.

Table 1. Data from the different populations examined in the present study.

Data Source	Latitude (DD)	Longitude (DD)	Year	Area	Site	Site in Fig.2	Depth (m)	% affected
This study	43.07530556°N	5.74944444°E	2022	Cap Sicie (France)	Les Moulinières	1	20	90
This study	43.07444444°N	5.74944444°E	2022	Cap Sicie (France)	Pierre de la Chaise	1	20	86
T-MEDNet	44.02377778°N	9.85044444°E	2004	La Spezia (Italy)	Tinetto Island	2	21	78.0
T-MEDNet	44.02377778°N	9.85044444°E	2006	La Spezia (Italy)	Tinetto Island	2	21	84.0
This study	44.17072222°N	9.58541667°E	2022	Levanto (Italy)	Levanto	3	25	30.3
This study	43.46269444°N	10.34163889°E	2022	Livorno (Italy)	Calafuria shallow	4	25	54.1
This study	44.25002778°N	9.40486111°E	2022	Sestri Levante (Italy)	Manara	5	25	29.2
This study	43.02652778°N	6.09711111°E	2022	Giens (France)	Escampobarieu	6	25	82.8
This study	40.68130556°N	14.43163889°E	2022	Gulf of Napoli (Italy)	Banco Santa Croce	7	25	28.0
This study	43.04725000°N	5.86736111°E	2022	Cap Sicie (France)	Pierre Benoît	8	25	85
This study	43.04725000°N	5.86736111°E	2022	Cap Sicie (France)	Pierre Benoît	8	25	95
This study	43.04311111°N	5.86336111°E	2022	Cap Sicie (France)	Sec des Pêcheurs	9	25	77
This study	38.26047222°N	15.71858333°E	2022	Sicily (Italy)	La Montagna (Scilla)	10	25	60
T-MEDNet	44.38900000°N	8.92669444°E	1993	Genova (Italy)	Gulf of Genova	11	27.5	80.0
This study	43.04755556°N	5.86930556°E	2022	Cap Sicie (France)	Arroyo	12	28	40.4
This study	43.04725000°N	5.86736111°E	2022	Cap Sicie (France)	Pierre Benoît	8	29	14
T-MEDNet	43.49330556°N	15.96916667°E	2009	Croatia	Cutin Mali	13	34	10
This study	44.17072222°N	9.58541667°E	2022	Levanto (Italy)	Levanto	14	35	13.2
This study	43.46269444°N	10.34163889°E	2022	Livorno (Italy)	Calafuria deep	4	35	17.4
This study	44.25002778°N	9.40486111°E	2022	Sestri Levante (Italy)	Manara	5	35	9.0
This study	43.02652778°N	6.09711111°E	2022	Giens (France)	Escampobarieu	6	35	12.0
This study	42.65197222°N	8.94188889°E	2022	Corsica (France)	Ille Rousse	15	35	7.6
This study	43.04311111°N	5.86336111°E	2022	Cap_Sicie (France)	Sec des Pêcheurs	9	35	49
This study	43.04311111°N	5.86336111°E	2022	Cap_Sicie (France)	Sec des Pêcheurs	9	35	10
T-MEDNet	44.97025000°N	14.82122222°E	2009	Croatia	Cape Sokol	16	35	40.0
T-MEDNet	44.70416667°N	14.73388889°E	2009	Croatia	Lun	17	35	10
T-MEDNet	43.51330556°N	15.91861111°E	2009	Croatia	Hrid Mulo	18	35	10
T-MEDNet	43.51055556°N	15.94222222°E	2009	Croatia	Smokvica Vela	19	35	10
T-MEDNet	43.49330556°N	15.96916667°E	2009	Croatia	Planka	20	35	10
This study	44.26180556°N	8.46363889°E	2022	Vado Ligure (Italy)	Magazzino	21	40	7.9

Continued

Table 1 continued

Data Source	Latitude (DD)	Longitude (DD)	Year	Area	Site	Site in Fig.2	Depth (m)	% affected
This study	43.04311111°N	5.86336111°E	2022	Cap_Sicie (France)	Sec des Pêcheurs	9	40	10
T-MEDNet	39.52000000°N	2.70000000°E	2003	Spain	Balearic Islands	22	40	14.9
This study	42.99319444°N	6.27436111°E	2022	Porquerolles (France)	Donator	23	45	7.3
This study	42.65197222°N	8.94188889°E	2022	Corsica (France)	Ille Rousse	15	45	9.0
This study	43.40463889°N	6.85238889°E	2022	St Rapahael (France)	Dramont	24	50	1.9
This study	40.68130556°N	14.43163889°E	2022	Gulf of Napoli (Italy)	Banco Santa Croce	7	50	4.8
This study	43.43938889°N	6.92205556°E	2022	St Rapahael (France)	Roux	25	60	1.1
This study	40.56863889°N	14.28675000°E	2022	Gulf of Napoli (Italy)	Secchitella	26	60	3.8
This study	38.26047222°N	15.71858333°E	2022	Sicily (Italy)	La Montagna (Scilla)	10	60	8
T-MEDNet	36.82444444°N	12.01444444°E	2011	Sicily (Italy)	Cala Levante-Pantelleria	27	60	10
This study	43.16602778°N	6.59519444°E	2022	Cavalaire (France)	Togo	28	65	5.2
This study	39.13630556°N	9.39286111°E	2022	Sardinia (Italy)	Loredan	29	65	1.7
This study	39.10627778°N	9.59930556°E	2022	Sardinia (Italy)	San Marco	30	90	1.5

Bearing in mind that mass mortality events have been mainly associated with MHWs, and given the well-established decrease in temperature with depth, even within the context of a thermal anomaly regime (Juza *et al.*, 2022; Dayan *et al.*, 2023), deeper populations should be preserved (but see Neal *et al.*, 2014; Venegas *et al.*, 2019). Nevertheless, it is important to note that both the intensity and duration of MHWs-related thermal anomalies has increased and reached deeper waters over the past decade (Dayan *et al.*, 2023).

In this study, we quantify the health status of *P. clavata* populations down to 90 m depth in the Mediterranean Sea (Fig. 1) and investigate the depth refuge hypothesis, i.e., we assess the maximum depths to which this species was impacted by MHWs. Our study attempts to elucidate the potential role of the mesophotic zone as a refuge for this species severely impacted by the effects of climate change.

Materials and Methods

We conducted mortality surveys on 43 populations located outside marine protected areas (MPAs) along the Western Mediterranean Sea, focusing on depths ranging from 20 to 90 m (Table 1; Fig. 1). A total of 32 surveys were collected by SCUBA diving by the authors of the present study between October 2022 and March 2023, while the remaining 11 surveys were extracted from the MME-T-MEDNet database (Cerrano *et al.*, 2000; Cupido *et al.*, 2008; Garrabou *et al.*, 2009, 2019).

A standardized protocol for underwater data collection was applied, based on a rapid assessment method to determine the health status of the colonies (Garrabou *et al.*, 2019) (t-mednet.org/MME). Figure 2 shows colonies of *P. clavata* both affected by mortality event (shallow populations; Fig. 2A) and in good health status (deep populations; Fig. 2B) for comparison. Approximately 100 adult colonies (> 15 cm in height) were inspected at each location and scored as affected when more than 10% of the colony presented necrosis. In affected colonies, we distinguished between the occurrence of a denuded axis (AD), indicating a very recent impact, axis with epibiosis (AE), indicating that the necrosis could be due to a past mortality event, and the presence of both denuded axis and epibiosis (AED). Since the MME-T-MEDNet dataset on mortality percentages only refers to recent mortality, we considered as affected all those colonies presenting > 10% of the denuded axis (AD + AED).

Two exclusion criteria were applied to the MME-T-MEDNet dataset:

- 1) Protection effect: According to Zentner *et al.* (2023), MPA can mitigate the effects of climate change on *P. clavata* populations by reducing sources of mortality other than water temperature increase (e.g., SCUBA divers' frequentation). Therefore, only data from populations dwelling outside MPAs were included, to avoid confounding effects between depth and protection status.
- 2) Accuracy of depth recording: In the MME-T-MED-

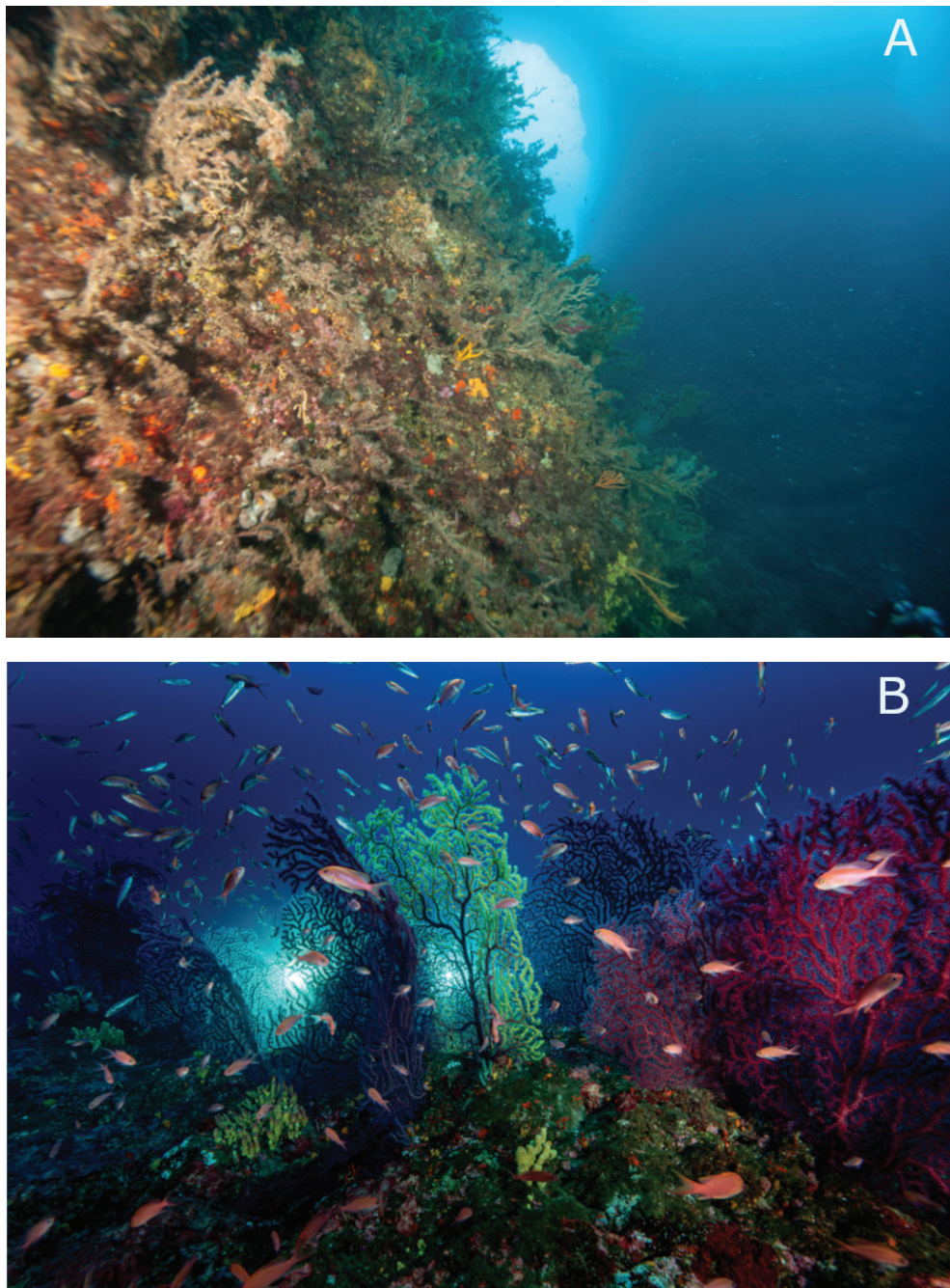


Fig. 2: Deep refuge. A: a shallow population (< 40 m) with a high percentage of affected colonies (Photo credits: T. Estaque). B: a deep population (> 40 m) with no signs of mortality (Photo credits: A. Rosenfeld, UNESCO and 1 Ocean Foundation).

Net dataset, the population survey depth is not given precisely but as a depth range with an upper and a lower bound. As we intend to establish a relationship between depth and percentage of affected colonies, accurate depth information is crucial. Therefore, we selected only those populations where the depth range (uncertainty) was less than 10 m and used the mean between the upper and lower bounds.

The relationship between depth and percentage of affected colonies was investigated following a three-steps procedure. Firstly, we selected the model that best described the relationship between the two variables using an Akaike information criterion (AIC) approach (Akaike, 1978). Testing linear, power, and exponential models we retained the power model as it resulted in the lowest

AIC value. Secondly, the strength of the non-linear relationship between the two variables was quantified by the Spearman correlation coefficient. Thirdly, the functional relationship between the variables was described using a non-linear least square (NLS) regression model (using the R software, function “stat_smooth”). A Student t-test was applied to assess the difference in mean Boxplot representing the percentage of affected colonies in shallow (< 40 m depth) and deep (> 40 m depth) *P. clavata* populations. -homogeneity of variance, a log transformation of the data was applied. The choice of the depth threshold of 40 m was based on previous studies suggesting the absence of MHWs-related impact below this depth (Arnoux *et al.*, 1992; Linares *et al.*, 2005; Pérez-Portela *et al.*, 2016; Estaque *et al.*, 2023).

Results

The percentage of affected colonies in the 43 populations of the present study varied from 95% at Cap Sicié (25 m depth; La Seyne sur mer, France) to 1.1% at Cap Roux (60 m depth; St. Raphael, France; Table 1; Fig. 2).

A significant negative correlation between the percentage of affected colonies and depth was found ($\rho = -0.89$, $p < 0.001$).

The relationship between depth and percentage of affected colonies was best described by a decreasing power curve with exponent -3 (Fig. 3). The regression model showed that a significant and high proportion of mortality variability was explained by depth ($F_{(1,41)} = 147.25$, $p\text{-value} < 0.001$, $R^2 = 0.78$).

A difference in the mean values of the percentage of affected colonies was found between shallow (< 40 m depth; $42.1\% \pm 32.1$ SD) and deep populations (> 40 m depth; $6.2\% \pm 4.1$ SD) (Fig. 4; $t\text{-test}$: $t = -6.23$ $df = 41$, $p\text{-value} < 0.001$).

Discussion

According to the “deep reef refugia” hypothesis (DRRH), marine populations inhabiting deeper zones should be less impacted by climate change-associated disturbances, particularly those stemming from rising seawater temperatures. The pivotal role of depth in mitigating this impact is widely accepted (Glynn, 1996; Bongaerts *et al.*, 2010; Bridge *et al.*, 2014; Bongaerts & Smith, 2019).

Our findings suggest that the mesophotic zone below 40 m depth might serve as a refuge for the Mediterranean *P. clavata*, providing shelter from the effects of marine heatwaves (MHWs), at least until 2022. Although the dataset used in this study is the most comprehensive and covers the broadest bathymetric range to date, our knowl-

edge of this species’ spatial distribution and conservation status beyond 40 m remains scant because of the limited number of available observations below this depth (but see Linares *et al.*, 2022 <https://cormednet.medrecovery.org/>). This dearth of knowledge impedes the precise delineation of bathymetric boundaries for such mortalities, and hampers the establishment of consistent knowledge base to forecast the impact of MHWs on *P. clavata* populations. The present study simultaneously provides empirical evidence of a higher survival of this gorgonian at greater depth and strongly advocates for further investigations of the mesophotic zone to chart potential areas of deep refuge for *P. clavata*. In this context, the recent democratization of technological solutions to explore the marine environment (e.g., closed circuit scuba diving, cabled observatories, remotely operated vehicles) provides a tangible avenue for deploying novel monitoring programs.

It should be noted that the observed refuge effect is likely to change in the near future, due to the expected increase in MHWs frequency and severity (Dayan *et al.*, 2023), which means that these events will extend their influence also in deeper, currently sheltered, areas. Historically, MHW events have demonstrated moderate impacts on gorgonian populations down to 65 m in Italy (Cerrano *et al.*, 2000; Pérez *et al.*, 2000). Furthermore, thermal tolerances described by previous investigations of shallow populations may not be directly transferable to their deeper counterparts (e.g., Crisci *et al.*, 2017; Gómez-Gras *et al.*, 2022). A decline in shallow populations could augment the level of isolation of deep populations, possibly resulting in higher sensitivity to thermal anomalies due to local adaptations to colder and more thermally stable conditions, thus elevating population vulnerability (Gómez-Gras *et al.*, 2022). In the specific case of *P. clavata*, thermotolerance in shallow populations (< 40 m depth) does not appear to be influenced by depth, and substantial inter-individual variability in responses to

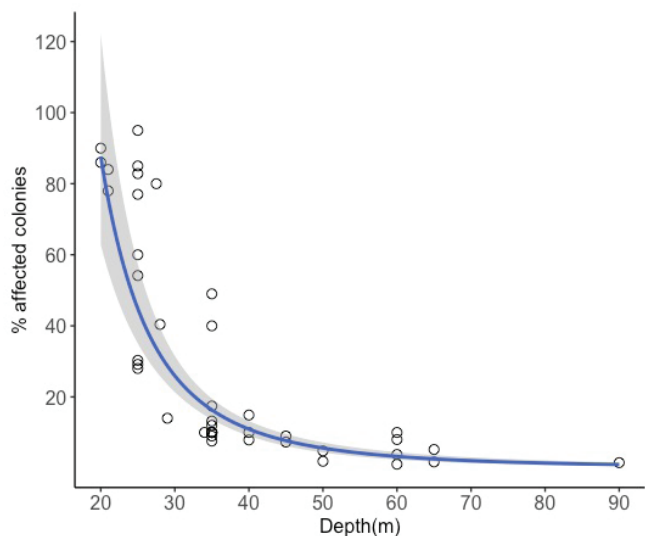


Fig. 3: Decreasing power curve describing the relationship between mortality (% of affected colonies) and depth. The grey area represents the 95% confidence interval.

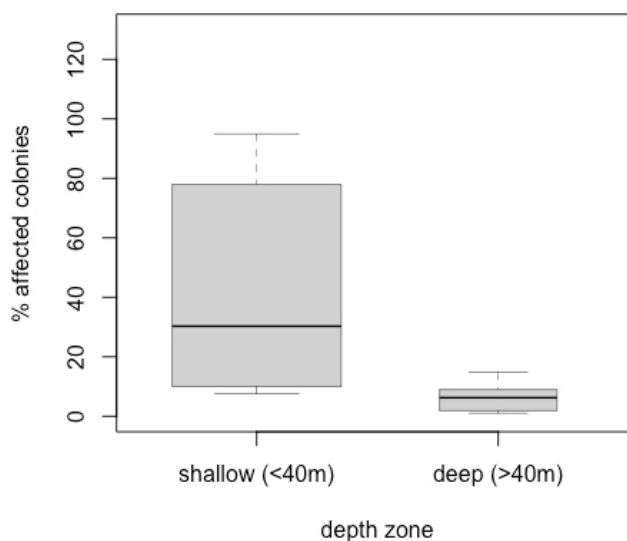


Fig. 4: Boxplot representing the percentage of affected colonies in shallow (< 40 m depth) and deep (> 40 m depth) *P. clavata* populations.

thermal stress implies the potential for local adaptation among isolated populations (Gómez-Gras *et al.*, 2022). However, adaptation processes in deep populations have not been investigated so far. A crucial step towards understanding the potential resilience of species with broad bathymetric ranges involves assessing genetic connectivity between shallow and deep populations. Regrettably, such studies on connectivity across varying depths are scarce (Mokhtar-Jamaï *et al.*, 2011; Pilczynska *et al.*, 2016), and only one study exists that has focused on deep populations (Pérez-Portela *et al.*, 2016). On the other hand, studies on vertical connectivity between *P. clavata* populations indicate limited gene flow, typically constrained to distances of approximately 100 m (Mokhtar-Jamaï *et al.*, 2011). Nonetheless, a study documenting the potential recovery of an impacted *P. clavata* population in the Ligurian Sea, demonstrated gene flow spanning over 30 km (Padrón *et al.*, 2018b).

Given the rather somber forecast for shallow populations (Gómez-Gras *et al.*, 2021), their survival in deep refuges depends on the degree of demographic connectivity between deep populations. In light of these considerations, the data presented here and the gaps in our understanding regarding the demography and physiology of deep populations underscore the compelling need to study mesophotic zones and elucidate connectivity among endangered populations through genetic connectivity and larval dispersal investigations (Sciascia *et al.*, 2022). This emerging field in conservation ecology will need a more precise delineation of thermal seascapes across bathymetric gradients and their influence on species distributions. Only through this enhanced understanding, the projection models of MHWs (e.g., Martínez *et al.*, 2023) can be effectively linked to the potential occurrence of disease outbreaks in deep populations, thereby enabling the forecast of the future trajectory of deep refuges.

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