Nanofiltration and selective crystallisation for the recovery of bio-based fertilizing products from seawater desalination brine

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Abstract. The majority of companies that operate desalination units produce significant quantities of brine, a hypersaline product whose disposal in the sea contributes to the degradation of local fauna and flora. The proposed framework for the valorisation of seawater brine in its entity and recovery of high purity water unveils a sustainable zero liquid discharge strategy consisting of two precipitation steps for the recovery of Mg (OH)2 and CaCO3. After pH conditioning, the brine (without Mg²⁺ and Ca²⁺) is led to nanofiltration unit for Na₂SO₄ separation from the NaCl - KCl rich stream. Monovalent salt stream is further concentrated by multiple effect distillation evaporator and crystallizer and KCl is separated from NaCl by flotation using sodium dodecyl sulfate as floating agent. The outcomes of this exploratory effort validate that the process upscaling is feasible. The aforementioned micronutrients have readily been tested for their part in enhancing the agronomic performance of fertilising products, so they possess market potential to support local/national/European fertilisation purposes. Last but not least, such a nutrient recovery approach can ensure capital recovery within the expected time limitations of an investment and in some cases offset completely the treatment cost.

Keywords: nutrient recovery, seawater desalination brine.

Table 1. List of abbreviations

Abbreviation	Name	Abbreviation	Name
BBFs	Bio-based fertilisers	N	nitrogen
Ca	calcium	NR	nutrient recovery
$CaCO_3$	calcium carbonate	С	organic
KCl	potassium chloride	P	phosphorus
$Mg (OH)_2$	magnesium hydroxide	K	potassium
MED	multiple effect distillation	RO	reverse osmosis
NF	nanofiltration	WW	waste water

1 Introduction

Climate change, climate immigrants, and the global population growth challenge the future of crop production [1] by intensification [2] and compliance with strict environmental rules at a local/national/international level [3]. This trend demands that mineral-based fertilizers be replaced because of variations in nutrient availability across different regions, intermittent restrictions on nutrient transfer, and growing awareness of ore depletion [4].

To both combat these challenges and make agriculture sustainable and greener, several organisations are currently committed to the application of nutrient recovery (NR) strategies for the production of fertilising products [4]. The nutrients in solid waste and waste water (WW) streams are considered as raw material for obtaining Bio-based fertilisers (BBFs) to replace mineral-based fertilisers. The feedstock [5] can be anything from solid (all ashes with phosphorus (P) content >7%, precipitated phosphate salts (struvite [6] (after calcining to remove organic contaminants, with ammonia recovery), calcium phosphate), food waste, food industrial by-products (food processing wastes) or bone meal, secondary materials e.g. wood chips, organic fibres, biomass ash, sewage sludge digestates, ashes from sewage sludge incineration or liquid waste streams, thought attempts have shifted towards the valorisation of nutrient-based, commercially significant products found in WW streams. The WW feedstock candidates for NR practices can be liquors with a high ammonium concentration, and food industrial WW, any type of industrial WW (paper, food processing, bottle water and cosmetics), urban WW, as well as brine, the target of the presented work.

Currently, NR technologies involve the recovery of both P and nitrogen (N) and there are cases in which micronutrients are also recovered from the input streams. As an example of crystallization as calcium silicate-hydrate, mono or diammonium phosphate can be recovered by precipitation with phosphoric acid by reacting with ammonia gas [7]. Other examples of process technologies can include microbial fuel cells [8] or any kind of hybrid systems of photoautotrophic and heterotrophic microalgal cultivation [9], nanofiltration (NF), multiple effect distillation (MED). In order to validate the market of recovered nutrients and the nutrient recovery technologies, SuSPhos® 'has built and validated a kilogram-scale prototype reactor and is now ready for the compilation of the industrial pilot' [10]. Using a hollow fibre liquid-liquid membrane contactor (HF-LLMC) is a promising technology capable of transforming recovered nitrogen into several ready-to-use fertilisers (ammonium salts). This technology is capable to produce ammonium salts in solution with a concentration like conventional fertilisers present in market [11].

Regarding the recovered products, they should be part in the fertilisers market either by replacing an existing product or by complementing its performance. In some countries, struvite from NR practices, already has an EU 2003/2003 fertiliser validation and an end-of-waste status [7]. The market share for organic and mineral fertilisers could be complemented by a BBF with an organic (C) and mineral (N, P) base. A substitute for nitrate fertilisers could be a BBF with a mineral base and a high N content. A BBF enriched in Ca/P with better properties (organic content) than mineral fertilisers of phosphate origin has the capacity to share the market for phosphate fertilisers. Another

example of WW-based NR product is Ammonium salts with plant growth promoting bacteria. These fertilisers can provide the inorganic nutrients required by the plant while the organic part can act as a bio-stimulant [12].

Any series of recovered micro-nutrients that are proven to complement plant growth can increase the yield of BBFs or replace micronutrient fertilisers on the market. A mixture of inorganic (N) and organic components could compete with products that are currently present in the market of organic and inorganic fertilisers. Hence, the relevance of the positioning in the market of this type of products is of utmost importance [13]. All in all, it has to be kept in mind that NR and the production of bio-based fertilisers, and WW-based fertilisers, in particular, are innovative practices [14]. The valorisation however of bio-based input streams (Urban WW, sewage sludge, food, Industrial WW and brine) into marketable fertilizing products or components is urgent and is guided and directed by accurate legislation for such products to enter the EU market. Still, the market for products from practices of recovery remains scarce [15].

Desalination of seawater is widely used to produce high-purity water to deal with water scarcity throughout Europe and the other continents. It is carried out in specialized plants using different techniques. The principal method for high-purity water production is reverse osmosis (RO). The application of this method results in fresh water and brine; a hypersaline by-product considered waste with a high concentration of minerals and metals [16]. Since seawater brine represents a potential alternative source for raw materials production and some salts are already being recovered, the development of suitable technologies for the selective recovery of a broad spectrum of salts as marketable products is imperative. Membrane-based separation processes, bioprocesses along with electrochemical and other hybrid processes represent technologies that have shown potential for metals recovery/removal [17]. The precipitation of minerals from brine reduces the fresh water production cost as well as minimises the brine disposal problem [18]. Despite their encouraging performance in lab-scale studies, these processes have not reached commercialisation yet, due to several associated challenges.

Current practice in countries using large-scale desalination plants is to reject brine back to the sea, leading to the degradation of local fauna and flora. However, further processing of brine could result in the recovery of economically prime important materials such as magnesium, calcium, and potassium salts and high-purity water. Brine treatment can contribute to the minimization of the environmental footprint of desalination plant [19].

Given its limited geological sources, Mg²⁺ has been qualified as a critical raw material by the European Union Commission [20]. Thus, its recovery from brine could be an alternative source. Another salt with an increasing consumption rate is KCl. It is estimated that under current consumption, world stocks of KCl will last until 2111. KCl is another salt that can be recovered from seawater desalination brine, as mentioned in the previous paragraph.

Due to its low cost and higher potassium (K) content than most other sources (50–52 % K, 60–63 % K₂O and 45–47 % Cl⁻), potassium chloride (KCl) is the most often used K fertilizer [21]. In addition to achieving a sustainable circular economy, recovery and reuse of potassium are necessary since its current consumption rate will soon lead to the depletion of its reserves.

The aim of this study is to investigate the process parameters for the optimization of Mg(OH)₂, limestone calcium carbonate CaCO₃ and KCl recovery in terms of valorization of brine as a wastewater stream from the desalination plants. The main targets are the recovery of high purity water and salts. This work is the predecessor of the design of a pilot system for the treatment of the waste stream produced from a RO desalination plant. These salts are economically prime important nutrient resource-based micro and macro nutrients to be valorised in/as Bio-based fertilisers. Potassium can be reused as a primary macronutrient, while Mg²⁺ and Ca²⁺ as micronutrients for the manufacturing of bio-based fertilizers [22, 23]. The parameters examined are reagents quantity and stirring time to optimize the purity and quantity of the recovered salts that are obtained as marketable ingredients of BBFs. The experiments are analyzed and organized through a 2² factorial design [24] in order to make a statistical evaluation of the tests and results.

2 Methodological approach

2.1 Brine sampling

Brine samples were collected during three different periods of time from the outlet of a small-scale desalination plant applying reverse osmosis (RO) of a hotel in Sounio, south-east Athens. The samples were stored at ambient conditions.

2.2 Brine characterisation

Analyses were performed according to standard methods (APHA). Every experimental measurement is preceded by the analysis of a brine sample, from the same day as the experiment. The results of seawater desalination brine applying reverse osmosis (RO) and seawater (inlet of RO) characterisation are presented in **Table 8**. There are variations depending on the local composition of brine, therefore the measured ion concentrations are likely to slightly vary in different brine samples.

2.2.1 Materials

Sodium hydroxide-NaOH(2M), calcium hydroxide-Ca(OH)_{2(s)}, nitric acid-HNO₃ 1.5%, sodium carbonate-Na₂CO_{3(s)}, potassium chloride-KCl, NaCl, sodium dodecyl sulfate-CH₃(CH₂)₁₁OSO₃Na and hexadecyl trimethyl ammonium bromide were all supplied by Sigma Aldrich (Merck).

2.2.2 Methods

A Spectroquant[®] UV-VIS spectrophotometer Pharo 300 was used to measure the chlorides (Cl⁻), the sulfate ions (SO₄²⁻), the nitrates (NO₃⁻) and the phosphates (PO₄³⁻).

The concentration of K⁺, Na⁺, Ca²⁺ and Mg²⁺ in the brine samples was measured using the Agilent Technologies[®] 200 Series AA 240 FS AA flame atomic absorption spectrometer (FAAS). All samples were analysed in triplicate. The setting for the examples was: wavelength 248.3 nm, slit width 0.2 nm, flame type: acetylene-air with a flow of 2.0 L/min with 13.5 L/min flow of air. The calibration curve, with deuterium

used for the background correction, ranges between 0.1–10 mg/L. The limit of detection (LOD) is 0.002 mg/L and 0.007 mg/L and the limit of quantification (LOQ) is 0.005 mg/L and 0.025 mg/L for the instrument and the method, respectively.

Bicarbonate ions (HCO³-) were measured with titration and neutralization up to pH=4.4.

Analyses on recovered dried samples with X-ray Diffraction (XRD) were performed using a Cu X-ray tube (λ =1.5418 Å) BRUKER D8-ADVANCE (twin/twin) diffractometer (40kV, 40 mA). The recorded range of angles was 20=10-60 ° and the recording speed 0.02 °/sec.

2.3 Brine treatment process

The first step of the brine treatment train is the precipitation of Mg²⁺ in the form of Mg(OH)₂. The addition of two reagents NaOH and Ca(OH)₂ for Mg²⁺ recovery is studied. Two series of experiments are carried out under different values of stirring time and quantity of reagent to recover the best combination regarding the highest quantity and purity of Mg²⁺ recovery. Na₂CO₃ is added to the brine after Mg²⁺ removal for CaCO₃ precipitation. As in the former experiments series, stirring time and quantity of the reagent are the two parameters examined to achieve maximum recovery of CaCO₃. After pH conditioning, the brine (without Mg²⁺ and Ca²⁺) is led to NF unit for Na₂SO₄ separation. The stream rich in monovalent ions such as Na⁺, K⁺, and Cl⁻ is further concentrated by a rotary evaporator for the reclamation of high-purity water. Then, KCl is separated from the mixed salt of NaCl-KCl using a flotation technique. Sodium dodecyl sulfate and hexadecyl trimethyl ammonium bromide are used as flotation agents. The efficiency of each floating agent is examined in terms of KCl recovery and purity. The bench-scale experiments of the proposed seawater brine treatment are described in Fig. 1.

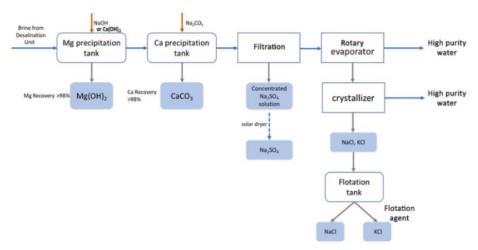


Fig. 1: Process flow diagram of the proposed seawater brine treatment

2.3.1 Mg²⁺ Removal

For the Mg²⁺ removal, samples of 200ml characterized brine are used. Chemical precipitation of Mg²⁺ is realised by using two different reagents. The first one is NaOH (2M) solution and the second one is solid Ca(OH)₂. Firstly, in every experimental series, the reagent is added under magnetic stirring. Different quantities of the reagent and different reaction times are examined. Sedimentation curves for magnesium hydroxide precipitation are also examined between 0-24 h. After the reaction, solid Mg(OH)₂ is separated through centrifugation (10min, 3500rpm) and then the liquid samples are collected. Liquid brine samples are diluted with nitric acid 1.5% and then are analyzed through AAS method to measure the final concentration of Mg²⁺ and calculate its removal percentage (%). Solid is washed out with deionized water, then collected in a crucible and left at 105 °C until dry. Finally, the dry solid is weighted, ground and analyzed through XRD method to determine its composition.

2.3.2 Ca²⁺ Removal

Solid sodium carbonate (Na₂CO₃) is added to brine samples without Mg²⁺ and the mixture is stirred. As a product of the reaction, a new solid is produced, CaCO₃, which is separated from the liquid solution through centrifugation (10min, 3500rpm). The liquid solution is filtered and then diluted with nitric acid 1,5% to be analyzed using AAS method. Through this analysis the concentration of Ca²⁺ is measured and the percentage (%) of Ca²⁺ removal is calculated. Solid CaCO₃ is washout out with deionized water. Finally, it is left at 105° C to dry and is analysed through XRD method.

After pH conditioning, the brine (without Mg²⁺ and Ca²⁺) is led to a rotary evaporator for water and salt recovery. Recovered salt is led to flotation for the separation of KCl from NaCl.

2.3.2.1 Factorial experimental design

Recovery of Mg^{2+} and Ca^{2+} is designed through a 2^2 factorial experiment of two parameters (reagent's quantity and stirring time). Experiments are realized under various conditions and combinations of the upper and lower levels of the parameters. The experiments with the conditions that refer to level 0 were repeated 4 times for statistical evaluation. The parameters selected are quantity of the reagent and different reaction times; a series of experiments are carried out under different conditions and combinations of the two parameters. X_1 refers to reagent's quantity and X_2 to stirring time. Stoichiometric quantity of the reagent is used as centre of the factorial design and based on the symmetry are defined the upper and lower extreme values of the levels. Specifically, at the upper level (+1) 110% of reagent's quantity is added with a stirring time of 45 min while at the lower (-1), 90% of the stoichiometric quantity is added with 15 min stirring time (**Table 2**).

Factorial parameters Factorial levels 0 unit 110% of 90% of stoig/ml Reagent's quantity X_1 stoichiometry stoichichiometry ometry 15 Stirring time X_2 30 45 min

Table 2: Parameters and levels of the factorial design

2.3.2.2 Factorial design for the recovery of Mg²⁺ from brine using NaOH (2M) as reagent

Experiments for the recovery of Mg²⁺, using NaOH (2M) as reagent are based on the chemical reaction (1):

$$Mg^{2+}_{(aq)} + 2 \text{ NaOH }_{(aq)} = Mg(OH)_{2 (s)} + 2 \text{ Na}^{+}_{(aq)} (1)$$

The initial sample is 200ml brine with 2,74g/L concentration in magnesium. Based on the above reaction, the stoichiometric quantity of NaOH (2M) for the precipitation of magnesium is 22.5ml; mixing time for this level is settled at 30'. Conducting all possible combinations of the levels of the two parameters (**Table 3**), resulted in 12 experiments in total. For the upper and lower level, 4 trials take place with two replicates for each level. As for the center, 4 replicates are realized for statistical validation.

Table 3. Factorial design for	r Mg ²⁺ recovery	from brine using NaOH	(2M) as reagent

Experiments A/A	Quantity NaOH (2M) (ml)	Mixing time (min)
1	20.25	15
2	20.25	15
3	24.75	45
4	24.75	45
5	20.25	45
6	20.25	45
7	24.75	15
8	24.75	15
9	22.50	30
10	22.50	30
11	22.50	30
12	22.50	30

2.3.2.3 Factorial design for the recovery of Mg²⁺ from brine using Ca(OH)₂ as reagent

Experiments for the recovery of Mg^{2+} , using $Ca(OH)_2$ as reagent are based on the chemical reaction (2):

$$Mg^{2+}_{(aq)} + 2 Ca(OH)_{2(s)} = Mg(OH)_{2(s)} + Ca^{2+}_{(aq)} (2)$$

Similarly to the latter process, 200ml of brine sample are used and based on the stoichiometry of the reaction, the quantity of Ca(OH)₂ for the Mg(OH)₂ precipitation is

1.356g under 30 min stirring time; these values are considered the center of the experimental design. Conducting all possible combinations of the levels of the two parameters (**Table 4**), resulted in 12 experiments.

Table 4. Factorial design for Mg²⁺ recovery from brine using Ca(OH)₂ as reagent

Experiments A/A	Quantity Ca(OH) ₂ (g)	Mixing time (min)
1	1.220	15
2	1.220	15
3	1.492	45
4	1.492	45
5	1.220	45
6	1.220	45
7	1.492	15
8	1.492	15
9	1.356	30
10	1.356	30
11	1.356	30
12	1.356	30

2.3.2.4 Factorial design for the recovery of Ca2+ from brine

Experiments for the recovery of Ca²⁺, using Na₂CO₃ as reagent are based on the chemical reaction (3):

$$Ca^{2+}_{(aq)} + Na_2CO_{3(s)} \rightarrow CaCO_{3(s)} + 2Na^{+}_{(aq)}$$
 (3)

Using 200ml of brine sample and based on the stoichiometry of the reaction (3), the quantity for the CaCO₃ precipitation is 0.48g Na₂CO₃ by stirring the mixture for 30'; these values are considered the center of the experimental design. Conducting all possible combinations of the levels of the two parameters (**Table 5**), resulted in 12 experiments.

30

Experiments A/A Quantity Na₂CO₃ (g) Mixing time (min) 1 0.43 15 2 0.43 15 3 45 0.53 4 0.53 45 5 0.43 45 6 0.43 45 7 0.53 15 8 0.53 15 9 30 0.4810 30 0.48 11 0.48 30

0.48

Table 5. Factorial design for Ca²⁺ recovery from brine

2.3.3 KCl recovery from brine

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After magnesium and calcium removal and pH conditioning from 10.5 to 7, the resulting brine is forwarded to a rotary evaporator for concentration and then to the oven for salt crystallization. KCl is more soluble at high temperatures than NaCl, so every 2 min, brine is separated from the crystallized salt. The separated brine is expected to be more enriched in KCl, upon full evaporation. Three types of salt resulted and are named A, B and C; C is expected to be the most KCl-rich, while A is the least KClrich fraction. Mixed (NaCl-KCl) salt is led to a flotation cell for the separation of KCl from NaCl. For this purpose, a saturated solution of KCl and NaCl is prepared by dissolving 342g KCl and 514g NaCl in 2L of deionized water. In every flotation trial, 10g of the mixed salts are added into 40 ml of the saturated solution and then Sodium Dodecyl Sulfate 1% is used as flotation agent. The mixture is stirred for 10 min and air is provided through a pipe to create froth; KCl crystals are attached on bubbles and then are collected. Finally, crystals are left to dry. A quantity of salts is diluted with nitric acid 1.5% for further qualitative analysis through AAS method. XRD analysis is performed for each kind of salt, to study their composition qualitatively. The results obtained are given in Fig. 2 to Fig. 4.

Commander Sample ID (Coupled TwoTheta/Theta)

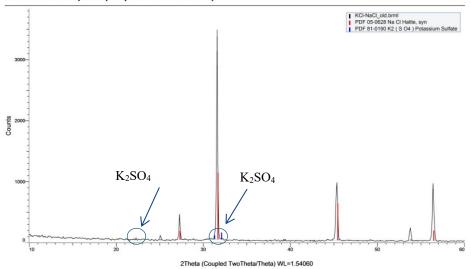


Fig. 2: Mixed salt A XRD analysis

Commander Sample ID (Coupled TwoTheta/Theta)

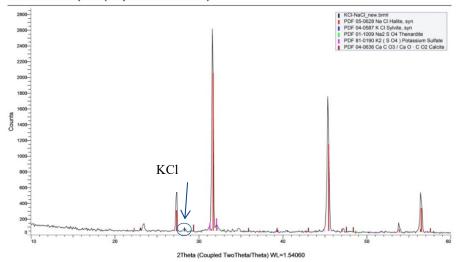


Fig. 3: Mixed salt B XRD analysis

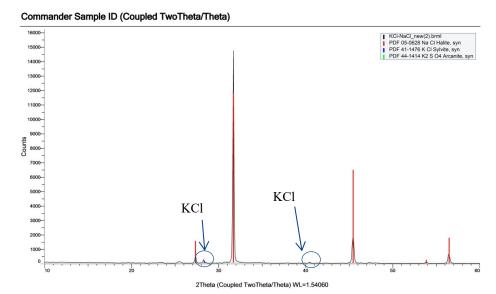


Fig. 4: Mixed salt C XRD analysis

Peaks corresponding to NaCl in the form of halite (red peaks) are visible in all three plots. In **Fig. 2**, which corresponds to salt A, some smaller peaks appear which are probably due to K₂SO₄ presence, while no peaks corresponding to KCl are found, due to the very low concentration of KCl in salt A. However, in Fig. 3 and in **Fig. 4** (mixed salt B and C) a small peak (blue in color) is visible due to the presence of KCl (sylvite). Other peaks probably correspond to the compounds K₂SO₄, Na₂SO₄ and CaCO₃.

To achieve a quantitative estimate of the composition of mixed salt, the analytical AAS is used (**Table 6**).

Table 6. Mixed salt analysis using AAS

K (mg/L)	Na (mg/L)	Ca (mg/L)	Mg (mg/L)	Fe (mg/L)
0.034	1.724	0.208	0.012	0.028

The concentration measured for potassium corresponds to 1.63 % KCl in salt B. There are also small concentrations of Mg²⁺ and Ca²⁺ which are probably not completely removed in the previous stages of the experiment. A total of 195 g of salt is collected where 52 g is salt A (less rich in K), 100 g salt B and 43 g salt C (richest in K). For the recovery of potassium by flotation method, 17 experiments are carried out where 10 g of salt are used for each one. The effect of the flotation agent addition is mainly studied keeping the amount of saturated KCl – NaCl solution constant; 40 mL are used in every experiment. In addition, tests are performed on how flotation time, mass of added salt and granulometry of the salt affect the recovered amount of KCl and its purity. **Table 7** shows the conditions for each experiment [25] and the recovered mass of solid.

Table 7.	Conditions	of the flotation	n experiments an	nd the rec	overed solid mass

Ex peri-	Pre- treatment time (min)	Flotation time (min)	Concentrated solution volume (mL)	Added salt (g)	Floating agent volume (mL)	Granulometry (μm)	Recovered solid (mg)
A1	5	5	40	10	5.0	-	360
A2	5	5	40	10	5.0	-	150
A3	5	5	40	10	2.5	-	200
A4	5	5	40	10	2.5	-	80
A5	5	5	40	10	1.0	-	70
B1	5	5	40	10	1.0	-	110
B2	5	5	40	10	5.0	-	180
В3	5	5	40	10	0.5	-	140
B4	5	5	40	10	0.5	-	80
B5	5	5	40	10	1.0	<180	60
B6	5	5	40	10	1.0	<180	60
В7	5	5	40	10	1.0	>180 & <1000	80
B8	5	5	40	10	1.0	>180 & <1000	100
C1	5	5	40	10	0.5	-	180
C2	5	5	40	10	0.5	-	130
C3	5	5	40	10	1.0	-	70
C4	5	5	40	10	1.0	-	100

As resulted from **Table 7**, recovered KCl quantity ranges from 60 mg to 200 mg with one single experiment providing 360 mg solid. According to the table, increasing the volume of the added flotation agent increases the recovered solid (**Fig. 5**) as more froth is collected [26].



Fig. 5. Recovered KCl crystals

3 Results and discussion

3.1 Feedstock characterization

Table 8 shows the concentration of main anions and cations contained in the three brine samples treated in this work and in seawater.

 Table 8. Ions concentration in seawater and brine

Ions concentration (mg/L)	Brine (1 st sample)	Brine 2 nd sample	Brine 3 rd sample	Seawater
K^+	813	874	906	380
Na	19674	17781	18230	8026
Ca^{2^+}	1552	1472	1523	450
Mg^{2+}	2736	2225	2200	1314
Cl ⁻	35400	44000	43800	27000
SO_4^{2-}	3820	5700	4730	4400
HCO ₃ -	238	107	221	140
PO_4^3	0.08	0.42	0.23	-
NO_3^-	0.01	1.10	0.80	-
TDS	6.85 g/100ml	7.43 g/100ml	7.48 g/ 100ml	3.45 g/ 100ml
Electrical conductivity	64.2 mS/cm ²	64.2 mS/cm ²	64.2 mS/cm ²	50.000 uS/cm
рН	7.77	7.70	7.80	7.50-8.40

3.2 Results from the factorial design

3.2.1 Mg²⁺ recovery with NaOH (2M) addition

According to **Table 9**, the maximum Mg^{2+} removal and minimum Ca^{2+} removal have occurred on experiments 5 and 6. In these cases, 90% of reagent's stoichiometric quantity is mixed with brine sample for 45 min.

Table 9. Mg(OH)₂ recovery through brine reaction with NaOH (2M)

N/N	Experiments A/A	Quantity NaOH (2M) (ml)	Mix- ing time (min)	%Mg ²⁺ removal	%Ca ²⁺ removal	Mg ²⁺ /Ca ²⁺	Recovered Mg(OH) ₂ mass (g)
N1	1	20.25	15	86.50	67.20	1.49	0.76
	2	20.25	15	86.70	59.10	1.69	0.87
N2	3	24.75	45	99.00	82.30	1.22	0.90
	4	24.75	45	99.90	70.80	1.41	1.10
N3	5	20.25	45	89.90	48.80	2.05	0.60
	6	20.25	45	89.70	53.60	1.89	1.07
N4	7	24.75	15	98.80	69.20	1.43	0.97
	8	24.75	15	100.00	65.90	1.52	0.83
			(Centre			
Cen-	9	22.25	30	92.00	54.60	1.68	0.84
tre	10	22.25	30	96.50	49.70	1.74	0.93
	11	22.25	30	95.50	50.50	1.79	0.85
	12	22.25	30	94.70	51.20	1.75	0.87
	13	Maximur moval based ome		100.00	0.00	100.00	1.276

Fig. 6 shows that experiments with maximum amount of reagent and maximum stirring time (N2) result in largest amount of salts recovery.



Fig. 6. Mass of recovered Mg(OH)₂ for every trial

Table 10. Mg(OH)₂ recovery through brine reaction with NaOH (2M)

A/A	Quantity of NaOH (2M) (ml)	Stirring time (min)
N1	20.25	15
N2	24.75	45
N3	20.25	45
N4	24.75	15
Centre	22.25	30
Maximum	Stoichiometr	ry

After processing the factorial experimental results (**Table 10**) [24, 27], the following model is developed, correlating the $Mg(OH)_2$ recovery from brine (Y, %) with the selected controlling parameters:

Y = -2.228 + 0.102 * (NaOH (2M) addition) + 0.077 * (stirring time) - 0.003 * (NaOH (2M) addition) * (stirring time) (4)

Equation 4 shows that all parameters are statistically important including their interaction. Furthermore, both mathematical models proved to be adequate according to the Fisher criterion [25, 28, 29, 30, 31].

3.2.2 Mg²⁺ recovery with Ca(OH)₂ addition

According to Table 11, maximum Mg removal has occurred with higher reagent dosage.

Table 11. Mg(OH)₂ recovery through brine reaction with Ca(OH)₂

N/N	Experiments No	Quantity Ca(OH) ₂ (g)	Mixing time (min)	%Mg re- moval	Recovered Mg(OH) ₂ mass (g)		
N1	1	1.220	15	86.88	0.936		
	2	1.220	15	87.46	0.881		
N2	3	1.492	45	99.98	1.652		
	4	1.492	45	99.62	1.716		
N3	5	1.220	45	86.91	1.145		
	6	1.220	45	86.64	1.195		
N4	7	1.492	15	99.98	1.030		
	8	1.492	15	99.18	0.940		
			Centre				
Centre	9	1.356	30	94.39	1.390		
	10	1.356	30	94.48	1.444		
	11	1.356	30	95.30	1.368		
	12	1.356	30	96.67	1.240		
	13	Maximum I	Mg ²⁺ removal	100.00	1.067		
based on stoichiometry							

Fig. 7 shows that experiments of N_2 trials provide the highest amount of salts recovery.

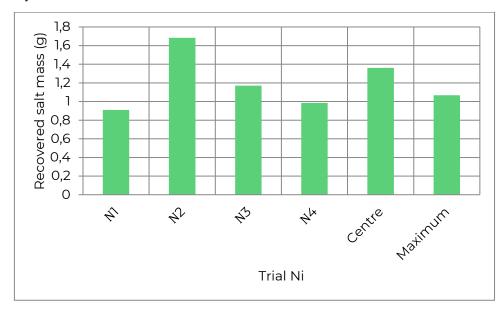


Fig. 7. Mass of recovered CaCO3 for every trial

First degree equation 5 shows that the most important parameter for Mg^{2+} recovery through $Ca(OH)_2$ addition is reagent's quantity, whereas reaction time is not considered as a factor that can affect the results.

 $Y = 0.295 + 0.47 * (Ca(OH)_2 addition) (5)$

3.2.3 Ca²⁺ recovery with Na₂CO₃ addition

Table 12 shows that all different experiments result in an average class rate of 99% CaCO₃ recovery.

N/N	Experiment No	Quantity Na ₂ CO ₃ (g)	Mixing time (min)	%Ca removal	Recovered CaCO ₃ mass (g)
N1	1	0.432	15	97.00	0.33
	2	0.432	15	98.00	0.35
N2	3	0.528	45	99.00	0.39
	4	0.528	45	98.00	0.34
N3	5	0.432	45	99.00	0.37
	6	0.432	45	99.00	0.38
N4	7	0.528	15	99.00	0.33
	8	0.528	15	99.00	0.37
		Cen	tre		
Centre	9	0.48	30	98.00	0.32
	10	0.48	30	99.00	0.35
	11	0.48	30	98.00	0.33
	12	0.48	30	99.00	0.37
	13	Maximum l	Mg ²⁺ removal	100.00	0.45
		based on sto	ichiometry		

Table 12. CaCO₃ recovery through brine reaction with Na₂CO₃

Equation (6) shows that Ca²⁺ removal increases both with Na₂CO₃ addition and stirring time. However, reagent's addition has a greater effect on Ca²⁺ removal because of its greater coefficient.

Y = 0.9372 + 0.0081 * (Na₂CO₃ addition) + 0.0002 * (stirring time) (6)

3.3 Nanofiltration simulation model

According to the experimental process flow diagram (**Fig. 1**) the brine after Mg²⁺ and Ca²⁺ removal will enter the NF unit. The occupation of the NF unit is necessary for the removal of SO₄²⁻ions from the seawater desalination brine. Regarding the NF process unit the separation efficiency was set as the main optimization parameter. Since only small volumes of brine (200ml and 50 L) have been treated during the lab-scale experiments, it was not possible to implement NF step. Thus, it was decided to run a simulation model for the treatment of 0.5 m³/day of brine (raw feed composition is presented in **Table 13**) using the water application value engine (WAVE) software.

	Raw Feed (mg/L)	Concentrate (mg/L)	Permeate (mg/L)
K	912.60	1,588	729.80
Na	21,974	37,701	17,718
Mg	11.60	30.57	6.46
Ca	17.40	45.98	9.66
SO_4^{2-}	3,416	15,046	271.30
TDS	58,688	103.304	46.616

Table 13. Concentrations of brine ions and TDS before and after NF process

After many runs with different types of NF membranes the most appropriate seems to be the membrane NF270-2540 in terms of SO₄-² removal. This membrane seems to be more appropriate (large pore diameter: 270 µm) because the stream is very dense and would create fouling and scaling issues [32]. The removal efficiency results in the brine salts & TDS removal percentages are described in **Table 14**. The simulation was performed under pH=7. For this type of membrane and aforementioned conditions no special comments were derived from the wave simulation model regarding fouling or scaling issues.

Table 14. NF % removal of brine ions & TDS

	% removal of major ions
K	20
Na	19
Mg	44
Ca	44
SO ₄ ² -	92
TDS	21
	•

To conclude with, the simulation of the NF process using membrane NF270-2540 implied a 92% removal (Table 14) of SO42- from the brine inlet. The resulting effluent streams will be used for the optimisation of the downstream processes.

4 Conclusions

Summing up, in the first part of the experimental process, the salts are recovered by precipitation as $Mg(OH)_2$ and $CaCO_3$ after the addition of NaOH or $Ca(OH)_2$ and Na_2CO_3 . Then, the divalent $SO_4{}^2$ are separated from the brine using nanofiltration (NF). The next part of the process targets brine concentration and the recovery of a mixed salt of NaCl and KCl applying thermal crystallization. In the final stage, KCl is collected from the mixed salt by flotation.

The main achievement of this work is the proof that macro and micronutrients can be recovered from the brine produced by an RO desalination plant. The acquisition of such end-products adds high value to the existing desalination plants, contributing to the alignment with EU Water Framework Directive, Circular Economy Package and sustainability in the agricultural sector. Furthermore, this work contributes to the definition of the main design of a system parameters and their values.

To achieve the recovery of water and salts from RO desalination brine next steps should be followed:

- o Mg(OH)₂ recovery (precipitation)
- o CaCO₃ recovery (precipitation)
- \circ Separation of the divalent SO_4^2 from the brine (nanofiltration).
- o Brine concentration (water evaporation)
- o Recovery of a mixed salt of NaCl and KCl
- o Separation of KCl from the mixed NaCl-KCl salt (flotation).

A sustainable brine treatment leads to avoidance of its disposal in the sea and decreases its impact on the environment and aquatic life. Also, recovery of nutrients from brine could reduce the CO₂ emissions in comparison with the conventional methods of nutrients production. Moreover, this technology facilitates compliance with proposed regional/EU-27 regulations towards replacing the production of fossil-based fertilizers. The introduced seawater brine technology framework outlines a sustainable, competitive zero liquid discharge approach to completely convert brine into reclaimed water and salts of commercial value. Last but not least, the results of this preliminary work justify the feasibility of the process upscaling.

5 References

- 1. Hincks S., Carter J., Connelly A., A new typology of climate change risk for European cities and regions: Principles and applications, Global Environmental Change 83, 102767 (2023) https://doi.org/10.1016/j.gloenvcha.2023.102767.
- Knapp M., Teder T., Lukas V., Štrobl M., Knappová J., Landis D.A., González E., Ecologically-Informed Precision Conservation: A framework for increasing biodiversity in intensively managed agricultural landscapes with minimal sacrifice in crop production, Biological Conservation 288, 110343 (2023) https://doi.org/10.1016/j.biocon.2023.110343.
- 3. Manghi M.C., Masiol M., Calzavara R., Graziano P.L., Peruzzi E., Pavoni B., The use of phosphonates in agriculture. Chemical, biological properties and legislative issues, Chemosphere, 283, 131187 (2021) https://doi.org/10.1016/j.chemosphere.2021.131187.
- 4. Egan A., Saju A., Sigurnjak I., Meers E., Power N., What are the desired properties of recycling-derived fertilisers from an end-user perspective?, Cleaner and Responsible Consumption 5, 100057 (2022) https://doi.org/10.1016/j.clrc.2022.100057.
- Ersahin M.E., Cicekalan B., Cengiz A.I., Zhang X., Ozgun H.: Nutrient recovery from municipal solid waste leachate in the scope of circular economy: Recent developments and future perspectives, Journal of Environmental Management, 335, 117518, (2023) https://doi.org/10.1016/j.jenvman.2023.117518.
- Achilleos, P., Roberts, K., & Williams, I., Struvite precipitation within wastewater treatment: A problem or a circular economy opportunity? Heliyon, e09862 (2022) https://doi.org/10.1016/j.heliyon.2022.e09862.

- 7. Luo Q., Xiang Y., Yang Q., Te Liang T., Xie Y.: Molecular simulation of calcium-silicate-hydrate and its applications: A comprehensive review, Construction and Building Materials 409, 134137, (2023) https://doi.org/10.1016/j.conbuildmat.2023.134137.
- 8. Deng L., Ngo H.-H., Guo W., Wang J., Zhang H.: Evaluation of a new sponge addition-microbial fuel cell system for removing nutrient from low C/N ratio wastewater, Chem. Eng. J. (2018), https://doi.org/10.1016/j.cej.2018.01.028.
- Yang Z., Pei H., Hou Q., Jiang L., Zhang L., Nie C.: Algal biofilm-assisted microbial fuel cell to enhance domestic wastewater treatment: nutrient, organics removal and bioenergy production, Chem. Eng. J. 332, 277–285 (2018) https://doi.org/10.1016/j.cej.2017.09.096.
- 10. Hušek M., Moško J., Pohořelý M.: Sewage sludge treatment methods and P-recovery possibilities: Current state-of-the-art, Journal of Environmental Management 315, 115090 (2022) https://doi.org/10.1016/j.jenvman.2022.115090.
- Cheraghdar A., Jazebizadeh M.H., Mansourizadeh A., Hojjati M.R., Lau W.J.: Liquidliquid extraction of phenolic wastewater through a surface modified poly(vinylidene fluoride-co-hexafluoropropylene) hollow fiber membrane contactor and process optimization, Journal of the Taiwan Institute of Chemical Engineers, 105111 (2023) https://doi.org/10.1016/j.jtice.2023.105111.
- Perera, M. K., Englehardt, J. D., & Dvorak, A. C.: Technologies for Recovering Nutrients from Wastewater: A Critical Review. Environmental Engineering Science, 511-529 (2019). https://doi.org/10.1089/ees.2018.0436.
- Tur-Cardona J., Bonnichsen O., Speelman S., Verspecht A., Carpentier L., Debruyne L., Buysse J.: Farmers' reasons to accept bio-based fertilizers: A choice experiment in seven. Journal of Cleaner Production, 406-416 (2018) https://doi.org/10.1016/j.jclepro.2018.06.172.
- Srivastava, V., Vaish, B., Singh, A., & Singh, R. Nutrient recovery from municipal waste stream: status and prospects. In Urban Ecology - Emerging Patterns and Social-Ecological Systems (pp. 265-297). Elsevier (2020) https://doi.org/10.1016/B978-0-12-820730-7.00015-X.
- Preisner, M., Smol, M., Horttanainen, M., Deviatkin, I., Havukainen, J., Klavins, M., Roosalu, K.: Indicators for resource recovery monitoring within the circular economy model implementation in the wastewater sector. Journal of Environmental Management, 114261 (2022) https://doi.org/10.1016/j.jenvman.2021.114261.
- Avramidi, M.; Spyropoulou, C.; Loizou, C.; Kyriazi, M.; Novakovic, J.; Moustakas, K.; Malamis, D.; Loizidou, M.: Adding Value to Reclaimed Water from Wastewater Treatment Plants: The Environmental Feasibility of a Minimal Liquid Discharge System for the Case Study of Larnaca. Sustainability 15, 14305 (2023) https://doi.org/10.3390/su151914305
- Panagopoulos, A.: Brine Management (Saline Water & Wastewater Effluents): Sustainable Utilization and Resource Recovery Strategy through Minimal and Zero Liquid Discharge (MLD & ZLD) Desalination Systems. Chem. Eng. Process.-Process Intensif., 176, 108944 (2022) https://doi.org/10.1016/j.cep.2022.108944.
- Frank, H.; Fussmann, K.E.; Rahav, E.; Zeev, E.B.: Chronic Effects of Brine Discharge from Large-Scale Seawater Reverse Osmosis Desalination Facilities on Benthic Bacteria. Water Res. 151, 478–487 (2019) https://doi.org/10.1016/j.watres.2018.12.046.
- Ogunbiyi, O., Saththasivam, J., Al-Masri, D., Manawi, Y., Lawler, J., Zhang, X., & Liu, Z.: Sustainable brine management from the perspectives of water, energy and mineral recovery: A comprehensive review. Desalination, 115055 (2021) https://doi.org/10.1016/j.desal.2021.115055.

- Charisiadis, C.: Zld booklet Lenntech, A quide to the basic conceptualization of the ZLD/MLD process design and the relative technologies involved. Lenntech- Water Treatment Solutions. Available at: https://www.lenntech.com/Data-sheets/ZLD-booklet-for-Lenntech-site-min-L.pdf, 67-70 (2018) https://doi.org/10.13140/RG.2.2.19645.31205.
- 21. Smol M.: Transition to circular economy in the fertiliser sector—analysis of recommended directions and end-users' perception of waste-based products in Poland, Energies 14, 10.3390/en14144312 (2021) https://doi.org/10.3390/en14144312.
- 22. Gorazda, K.; Tarko, B.; Wzorek, Z.; Kominko, H.; Nowak, A.K.; Kulczycka, J.; Henclik, A.; Smol, M.: Fertilisers production from ashes after sewage sludge combustion—A strategy towards sustainable development. Environ. Res. 154, 171–180 (2017) https://doi.org/10.1016/j.envres.2017.01.002.
- Ten Hoeve, M.; Bruun, S.; Naroznova, I.; Lemming, C.; Magid, J.; Jensen, L.S.;
 Scheutz, C.: Life cycle inventory modeling of phosphorus substitution, losses and crop uptake after land application of organic waste products. Int. J. Life Cycle Assess. 23, 1950–1965 (2018) https://doi.org/10.1007/s11367-017-1421-9.
- Alder, P., Markova, E.V., Granovsky, V.: The Design of Experiments to Find Optimal Conditions a Programmed Introduction to the Design of Experiments. Mir Publishers, Moscow (1975).
- 25. Allen, T.T.: Introduction to Engineering Statistics and Six Sigma. Springer, Colombus (2007).
- 26. Filippov, L. O., Filippova, I. v., Barres, O., Lyubimova, T. P., & Fattalov, O. O.: Intensification of the flotation separation of potash ore using ultrasound treatment. Minerals Engineering 171 (2021). https://doi.org/10.1016/j.mineng.2021.107092.
- 27. Cochran, W.G., Cox, G.M.: Experimental Designs. WIley, New York (1957).
- 28. Seidenfeld, T.: R. A. Fisher on the design of experiments and statistical estimation. 23–36 (1992) https://doi.org/10.1007/978-94-011-2856-8 2.
- 29. Pacheco-Torgal, F., Diamanti, M.V., Nazari, A., Granqvist, C.-G.: Nanotechnology in Eco-efficient Construction. Woodhead Publishing Limited, Swaston (2013).
- 30. Selvamuthu, D., Das, D.: Introduction to Statistical Methods, Design of Experiments and Statistical Quality Control. Springer, Singapore (2018).
- Cox, D.R., Reid, N.: The Theory of the Design of Experiments. Chapman & Hall CRC, New York (2000).
- 32. Kathijotes N., Panayiotou C.: Wastewater reuse for irrigation and seawater intrusion: evaluation of salinity effects on soils in Cyprus. J. Water Reuse Desalin. 3(4), 392–401 (2013) https://doi.org/10.2166/wrd.2013.072.